

1 **Risk Tradeoffs Associated with Traditional Food Advisories for Labrador Inuit**

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14

15 **Abstract**

16 The traditional Inuit diet includes wild birds, fish and marine mammals, which can contain high
17 concentrations of the neurotoxicant methylmercury (MeHg). Hydroelectric development may
18 increase MeHg concentrations in traditional foods. Consumption advisories are often used to
19 mitigate such risks and can result in reduced intake of traditional foods. Data from a dietary
20 survey, MeHg exposure assessment and risk analysis for individuals in three Inuit communities
21 in Labrador, Canada ($n = 1,145$) in 2014 indicate reducing traditional food intake is likely to
22 exacerbate deficiencies in n-3 polyunsaturated fatty acids and vitamins B12 and B2. Traditional
23 foods accounted for < 5% of per-capita calories but up to 70% of nutrients consumed. Although
24 consumption advisories could lower neurodevelopmental risks associated with an increase in
25 MeHg exposure (90th-percentile $\Delta IQ = -0.12$ vs. -0.34), they may lead to greater risks of
26 cardiovascular mortality (90th-percentile increase: +58% to +116% vs. +25%) and cancer
27 mortality (90th-percentile increase +2% to +4% vs. no increase). Conversely, greater
28 consumption of locally caught salmon mostly unaffected by hydroelectric flooding would lower
29 all these risks (90th-percentile $\Delta IQ = +0.4$; cardiovascular risk: -45% ; cancer risk: -1.4%). We
30 thus conclude that continued consumption of traditional foods is essential for Inuit health in these
31 communities.

32

33 **Keywords:** fish advisory, methylmercury, indigenous health, dietary transition, nutrition

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42 **Research ethics approval**

43 This work was carried out following the approval of the following research ethics authorities:

- 44 • Office of Human Research Administration, Harvard T.H. Chan School of Public Health
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- 46 • Newfoundland and Labrador Research Ethics Board (case 14.004)
- 47 • Nunatsiavut Government Health Research Ethics Authority

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51 **1. Introduction**

52 Traditional foods consumed by northern Inuit populations include locally caught fish,
53 birds and marine mammals. These foods are critical sources of micronutrients and high-quality
54 protein in regions that have limited access to other fresh foods. However, some traditional foods
55 contain elevated concentrations of contaminants such as methylmercury (MeHg) that biomagnify
56 in aquatic environments.¹ MeHg exposures among indigenous populations across Canada tend to
57 be higher than the national average due to relatively greater consumption of fish and marine
58 mammals.^{2,3} Fish consumption advisories are a common policy response to elevated
59 environmental levels of MeHg.⁴⁻⁷ However, even targeted food advisories can lead to decreased
60 overall consumption of seafood.^{8,9} In indigenous populations, some food consumption advisories
61 related to elevated MeHg levels have led to decreased overall consumption of seafood.^{10,11}

62 MeHg has a half-life of 50–70 days in the human body and thus dietary changes can alter
63 exposures over shorter timescales than many other hydrophobic organic pollutants such as
64 polychlorinated biphenyls (PCBs) that persist in the body for decades.^{12,13} Negative effects of
65 MeHg exposure on the brain of the developing fetus and young children have been shown to
66 persist into adulthood and this is the endpoint used by most regulatory agencies to establish risk
67 thresholds.¹⁴⁻¹⁶ An association between elevated prenatal MeHg exposures and
68 neurodevelopmental impairment has similarly been shown among Inuit children.^{17,18} MeHg
69 exposure has also been associated with cardiovascular health risks for adults.¹⁹⁻²²

70 Dietary advice to indigenous populations must balance the competing goals of
71 minimizing contaminant exposures with ensuring nutritional sufficiency.^{10,11,23} Thus,
72 quantitative studies quantifying the likely health risks or benefits from different dietary
73 interventions among indigenous populations are needed and are the focus of this work. Past
74 studies have examined risk tradeoffs between increased MeHg exposure and n-3 polyunsaturated

75 fatty acids (n-3 PUFAs) in seafood.²⁴⁻²⁷ Cardiovascular and neurodevelopmental benefits of n-3
76 PUFAs can offset negative impacts of MeHg exposure.²⁷ The World Health Organization
77 (WHO) and the Food and Agriculture Organization (FAO) have however recommended that
78 risk-benefit evaluations for seafood consumption consider a broader suite of nutrients, including
79 iron, zinc, and vitamins A, D, B2 and B12.²⁸

80 Dietary changes exert changes on health endpoints through diverse, competing
81 mechanisms. For example, red and processed meats increase cardiovascular risks by triggering
82 the formation of proatherosclerotic trimethylamines and raise colorectal cancer risks through
83 production of carcinogenic N-nitroso compounds in the gastrointestinal tract.²⁹⁻³¹ Conversely,
84 fruits, vegetables and nuts are rich in compounds that reduce cardiovascular risks by inhibiting
85 platelet aggregation and oxidation of arterial cholesterol (antioxidants and polyphenols),
86 mediating glucose homeostasis (fiber) and regulating blood pressure (potassium and
87 magnesium).³²⁻³⁶ There is strong evidence that vitamin D exerts a generalized cancer-protective
88 effect through the apoptotic and antiangiogenic properties of its metabolite 1,25(OH)₂D.³⁷⁻³⁹ The
89 Global Burden of Disease study synthesizes the most well-established causal relationships
90 between intake of various foods and nutrients and different cancers and cardiovascular
91 diseases.^{40, 41}

92 Here, we use results from a dietary survey and food subsidy data to better understand
93 consumption preferences and the nutritional importance of locally harvested foods for three Inuit
94 communities in the Lake Melville region of Labrador, Canada. In this region, local Inuit are
95 concerned about potential increases in MeHg exposure from traditional foods following
96 completion of a hydroelectric power facility upstream of their traditional hunting and fishing
97 territory. Flooding of hydroelectric reservoirs leads to a pulse in MeHg production in the

98 saturated soils, which can enter the overlying water and accumulate in fish and wildlife. In
99 previous work, we estimated increases in MeHg concentrations of traditional foods at their peak
100 due to the local hydroelectric development using a biogeochemical model.⁴² Modeled peak
101 increases in ranged from nine times 2014 levels for some freshwater species most affected by the
102 hydroelectric development to zero for offshore marine species.⁴²

103 In this study, we conduct a screening level analysis of potential risks associated with
104 higher MeHg concentrations in traditional foods. We compare the estimated magnitudes of
105 projected MeHg exposure risks to those attributable to changes in nutrient intake following a
106 dietary transition away from traditional foods using established dose-response relationships. We
107 conduct this analysis to better understand potential health implications of dietary consumption
108 advisories for indigenous populations more generally.

109 **2. Methods**

110 *2.1 Food frequency questionnaire and hair Hg analysis*

111 We developed a food frequency questionnaire (FFQ) to measure intake of traditional foods
112 and store-bought seafood among Labrador Inuit settled in three communities downstream from
113 the Churchill River: 1) Happy Valley – Goose Bay, 2) North West River, and 3) Rigolet (SI
114 Figure S1). We enrolled a total of 1,145 individuals, representing roughly 40% of the Inuit
115 population in these communities (SI Table S1). Responses were weighted by demographic
116 categories according to the 2011 Census to provide statistically relevant estimates for the whole
117 population (SI Table S2).⁴³ A team of 26 trained research assistants recruited from the Inuit
118 community administered the survey instrument.

119 The FFQ asked respondents to recall their intake of 64 traditional foods (locally caught
120 seafood, land mammals, birds, plants and berries) and 24 store-bought seafood types over 24-
121 hour, one-month and three-month recall periods. Foods measured by the FFQ are listed in SI

122 Tables S3 (locally caught seafood), S4 (other locally caught foods) and S5 (store bought
123 seafood). The FFQ was administered across three seasons to assess seasonal variability in
124 consumption preferences: Winter (March–April, $n = 231$), Spring (June–July, $n = 294$), and
125 Summer (August–September, $n = 1,054$). Enrollment was maximized for the Summer survey
126 period, which we use for risk calculations unless otherwise noted. Self-reported age, sex, height
127 and weight was included in the survey. All FFQ respondents in the Spring and Summer periods
128 were asked to provide hair samples. In total, 656 hair samples corresponding to 571 unique
129 individuals were collected (157 in Spring and 499 in Summer).

130 All work involving human subjects (recruitment, survey design, data analysis and
131 reporting) was reviewed and approved by the Office of Human Research Administration at the
132 Harvard T.H. Chan School of Public Health, the Newfoundland and Labrador Health Research
133 Ethics Authority and the Nunatsiavut Government Inuit Research Advisor. The Nunatsiavut
134 Government provided input on all research plans and has assumed responsibility for
135 disseminating research findings to community members and provincial and federal
136 policymakers.^{44,45} Additional information on the FFQ design and implementation, participant
137 recruitment and hair Hg analysis is provided in the SI.

138 2.2 Modeled MeHg exposures

139 We modeled MeHg exposures for all Inuit individuals in 2014 using dietary recall data for
140 the 88 traditional and store-bought foods and associated MeHg concentrations (modeled as
141 lognormal distributions). For locally caught foods, we directly measured MeHg concentrations in
142 22 species representing 81% of per-capita MeHg exposures from this category in 2014. All data
143 sources for MeHg concentrations were originally reported by Li et al.⁴⁶ et al. and Calder et al.⁴²
144 and are provided in SI Tables S3 (locally caught seafood), S4 (other locally caught foods) and S5
145 (store-bought seafood).

146 To correct for overreporting bias associated with species-specific recall, we scaled
147 reported species-specific intakes to match reported total consumption of three food categories
148 (local seafood, store-bought seafood and other locally caught foods) following Lincoln et al.⁴⁷
149 We probabilistically simulated hair mercury concentrations (10,000 Monte Carlo trials) for each
150 individual using the one-compartment model developed by the U.S. Environmental Protection
151 Agency's (U.S. EPA), lognormal distributions for MeHg concentrations in food items (Tables
152 S3–S5), and probabilistically distributed toxicokinetic parameters⁴⁸, following Li et al.¹²

153 The median ratio between measured hair Hg and simulated hair Hg was 0.96 in the
154 larger-scale Summer survey period, suggesting that there is very little bias in the dietary model
155 we developed (measured-to-modeled ratio close to 1). To ensure that population-wide MeHg
156 exposures are not overestimated by the dietary model, modeled seafood intake was scaled such
157 that simulated hair Hg matched measured Hg for all individuals with available hair Hg data. For
158 others, seafood intake was scaled by the median bias (0.96).

159 2.3 Dietary intake of other foods

160 We estimated consumption of market foods other than seafood that have high nutritional
161 content using sales data for the community of Rigolet for the same years as our dietary survey
162 (2014-2015).⁴⁹ Data on the edible supply of foods have been successfully used to estimate
163 dietary composition and caloric sufficiency in other populations.⁵⁰⁻⁵² Data were obtained from a
164 Canadian federal subsidy program (Nutrition North), which subsidizes 42 nutrient-dense and
165 perishable store-bought foods or food categories in remote communities.⁵³ We estimated
166 population-wide intake of nutrient-dense store-bought foods based on the magnitude of food
167 subsidies and by subtracting retail and consumer waste fractions⁵⁴ (SI Table S6). We assumed
168 the composition of market foods sold in Rigolet was similar for the other two communities
169 (Happy Valley-Goose Bay, Northwest River). Since traditional food intake was smaller in these

170 communities in 2014, we allowed for proportionally greater consumption of store-bought
171 nutrient-dense foods.

172 We used the relationship developed by Mifflin et al.⁵⁵ that has applied among indigenous
173 populations⁵⁶ to estimate total energy expenditure (E) of each Inuit individual based on self-
174 reported body mass, height, and age from survey data (eq. 1):

$$E = (9.99W + 6.25H - 4.92A + 166\alpha)\beta \quad [1]$$

175 where W is body mass (kg), H is height (cm), A is age (years) and $\alpha = 1$ and $\beta = 1.7$ for
176 men and $\alpha = 0$ and $\beta = 1.6$ for women. For each individual, the difference between estimated
177 energy expenditure and the caloric intake accounted for by locally caught traditional foods (data
178 from the FFQ) and nutrient-dense market foods (data from Nutrition North) was assumed to
179 correspond to comparatively nutrient-sparse foods such as potato chips and sweetened beverages.
180 There are very few foods that do not qualify for Nutrition North subsidies that have nutritional
181 value and, in other Inuit populations, they are not widely consumed.⁵⁷

182 2.4 Nutritional content of store bought foods

183 We used the Canadian Nutrient File⁵⁸ to estimate nutrient intake from store-bought
184 nutrient-dense foods and traditional foods. Government nutritional databases include traditional
185 and market foods consumed by Inuit populations and have been successfully used to estimate
186 population-wide nutrient intake for other indigenous groups (e.g., Quebec Inuit).⁵⁹ Five foods
187 with no data were matched to foods in the United States Department of Agriculture (USDA)
188 Nutrient Database⁶⁰. Nutrient contents were available for 90% of foods by calories. For other
189 foods, the average nutrient content of similar food categories was used. For this purpose, we
190 group all foods (traditional and store-bought together) according to the following food
191 categories: dairy, egg, fish liver, fish muscle, fish roe, fowl/poultry, fruit, grain, marine mammal,
192 processed meat, red meat, shellfish, terrestrial mammal and vegetables. The compiled database

193 of nutritional information for all foods studied here is included in the SI (Excel file: Appendix A,
194 SI Table S6). We consider that dietary supplementation has a negligible impact on population-
195 wide nutrient intake because dietary supplement tends to be rare among indigenous
196 populations.⁶¹⁻⁶³

197 2.5 Traditional food substitution scenarios

198 We developed five traditional food substitution scenarios to bound the possible range of
199 dietary changes that might occur in the future. These are: [1] Traditional foods are replaced by
200 nutrient-dense store-bought foods subsidized by the Nutrition North program (SI Figure S2). [2]
201 Traditional foods are replaced by processed meat as an alternative protein source. [3] Traditional
202 foods are replaced by vegetables. This scenario was used to provide a lower envelope of
203 nutritional risks. However, the Nunatsiavut Government identified this scenario as highly
204 unlikely given current consumption preferences. We nevertheless include it as a best-case
205 scenario for market-based food substitution. [4] Traditional foods are replaced by nutrient-sparse
206 foods such as snack foods. [5] Traditional foods high in MeHg are substituted with Atlantic
207 salmon. The Nunatsiavut Government identified Atlantic salmon as a preferred food item. We
208 use this scenario to investigate the health impacts of traditional diet adaptation instead of
209 replacement with store-bought foods. For all scenarios, caloric consumption is assumed to be
210 constant at the estimated 2014 values based on Eq. 1 above.

211 2.6 Screening-level risk assessment

212 We quantified the magnitudes of neurodevelopmental, cardiovascular and cancer risks
213 associated with the five dietary scenarios described above and compared them to those
214 associated with projected future MeHg exposures at 2014 diet. Projected future MeHg exposures
215 result from increased MeHg content in traditional foods as a result of upstream hydroelectric
216 development calculated by Calder et al.⁴². The probabilistic projections of future MeHg levels in

217 local traditional foods are included in SI Table S7. We do not consider cancer risks associated
218 with MeHg in traditional foods because the U.S. EPA classifies MeHg as a possible human
219 carcinogen, noting there is ‘no persuasive evidence’ for human carcinogenicity.¹⁶

220 Neurodevelopmental risks to children are expressed in terms of change in IQ and are
221 modeled by considering diets of women of childbearing age (16–49 following McDowell et
222 al.⁶⁴). We retain the confounder-adjusted dose-response functions summarized in Table 1. We
223 express cardiovascular and cancer risks as the relative risk (RR) of mortality compared to 2014,
224 calculated from changes in intake of various foods and nutrients. A RR of greater than 1.0
225 represents an increase in risk, and a RR of less than 1.0 represents a decrease in risk.

226 Cardiovascular and cancer risks are quantified as the product of individual confounder-
227 adjusted relationships following Fleming et al.⁶⁵. RRs are presented in the literature
228 corresponding to certain incremental doses. We assume these are proportional over the range of
229 incremental changes explored here and scale RRs presented in the literature to changes in food
230 substitution scenarios. We consider relative risks for cardiovascular and cancer deaths based on
231 diet for all individuals over 25 following Forouzanfar et al.⁶⁶. Dose-response functions for
232 cardiovascular and cancer mortality are expressed in terms of risks of more specific causes of
233 death (e.g., risk of cancer at certain sites). We consider the share of overall cardiovascular (Table
234 2) and cancer (Table 3) mortality represented by the outcome in each dose-response relationship
235 in order to calculate net impacts on the risk of total cardiovascular and cancer mortality. A
236 mathematical derivation (equations S1–S3) is presented in the SI.

237 Dietary dose-response functions for risk of cardiovascular and cancer death are based on
238 the Global Burden of Disease study.⁴⁰ We excluded benefits related to increased whole grain
239 consumption because we could not quantify the ratio of whole to processed grains in the baseline

240 diet and intake of whole grains is small in similar northern indigenous populations.^{61, 67} We
241 account for cardiovascular risks associated with consumption of red and processed meats and
242 benefits of fruits, nuts and vegetables as a function of intake of the whole food and thus do not
243 separately consider constituent nutrients in these foods (e.g., sodium in red meat). There is strong
244 evidence that vitamin D exerts a generalized cancer-protective effect, especially in northern
245 populations.³⁷⁻³⁹ We thus also consider a negative association between vitamin D and risk of
246 cancer mortality⁶⁸ following earlier work by Grant et al.⁶⁹ who calculated cancer mortality in
247 Canada attributable to vitamin D deficiency. This analysis is carried out using probabilistically
248 distributed parameters for relative risk of cancer and cardiovascular effects and for IQ gains and
249 decrements, allowing for explicit calculation of the uncertainties inherent to this analysis.
250

251 Table 1: Summary of dose-response relationships used for screening-level risk assessment of
 252 neurodevelopmental risk

Predictor	Outcome	Dose-response function ^a	Reference
MeHg	IQ points	1.07 per 0.5 g Hg g ⁻¹ hair (95% CI: 1.03–1.11)	Virtanen et al. ¹⁹
n-3 PUFAs	IQ points	mode = +1.3; min = 0.8; max = 1.8 per g DHA day ⁻¹	Cohen et al. ⁷⁰

253 ^a MeHg dose-response function is normal distribution and n-3 PUFA dose-response function is triangular
 254 distribution following the authors. Risks accrue to children born to mothers (females aged 16–49) with
 255 modeled intakes.

257 Table 2: Summary of lognormally distributed dose-response relationships used for screening-
 258 level risk assessment risk of cardiovascular mortality

Predictor	Outcome	Median (95% CI) relative risk per change (+) in intake	Reference
MeHg	SCD ^b	1.07 (1.03–1.11) per 0.5 g Hg g ⁻¹ hair	Virtanen et al. ¹⁹
n-3 PUFAs ^a	IHD ^c	0.866 (0.792–0.943) per 0.1 g day ⁻¹	Forouzanfar et al. ⁴⁰ , Chowdhury et al. ⁷¹
Fiber	IHD ^c	0.754 (0.678–0.831) per 20 g day ⁻¹	Threapleton et al. ²⁹ , Forouzanfar et al. ⁴⁰
Fruit	IHD ^c	0.867 (0.829–0.962) per 100 g day ⁻¹	Wang et al. ³² , Forouzanfar et al. ⁴⁰
Fruit	IS ^d	0.719 (0.604–0.8401) per 100 g day ⁻¹ (b)	Forouzanfar et al. ⁴⁰ , Hu et al. ⁷²
Fruit	HS ^e	0.868 (0.661–0.762) per 100 g day ⁻¹	Forouzanfar et al. ⁴⁰ , Hu et al. ⁷²
Nuts	IHD ^c	0.944 (0.845–0.914) per 4.05 g day ⁻¹ (b)	Forouzanfar et al. ⁴⁰ , Afshin et al. ⁷³
Processed meat	IHD ^c	1.603 (1.022–2.271) per 50 g day ⁻¹	Forouzanfar et al. ⁴⁰ , Micha et al. ⁷⁴
Trans fatty acids	IHD ^c	1.414 (1.281–1.567) per 2% energy intake ^f	Forouzanfar et al. ⁴⁰ , Mozaffarian and Clarke ⁷⁵
Vegetables	IHD ^c	0.96 (0.93–0.99) per 106 g day ⁻¹ (g)	Wang et al. ³²
Vegetables	IS ^d and HS ^e	0.89 (0.81–0.98) per 200 g day ⁻¹ (h)	Hu et al. ⁷²

259 ^a As DHA + EPA.

260 ^b Sudden cardiac death represents 27.4% of total cardiovascular mortality CDC ⁷⁶.

261 ^c Ischemic heart disease represents 70% of total cardiovascular mortality CDC ⁷⁶.

262 ^d Ischemic stroke represents 13% of total cardiovascular mortality CDC ⁷⁶.

263 ^e Hemorrhagic stroke represents 6% of total cardiovascular mortality CDC ⁷⁶.

264 ^f Total daily energy intake calculated from Mifflin et al.⁵⁵ as described in methods and assuming 9 kcal g⁻¹
 265 trans fatty acids ⁷⁷.

266

267 Table 3: Summary of lognormally distributed dose-response relationships used for screening-
 268 level risk assessment of risk of cancer mortality

Predictor	Cancer site (mortality)	Median (95% CI) relative risk per change (+) in intake	Reference
Calcium	Colon/rectum ^a	0.729 (0.831–0.963) per 1 g day ⁻¹	Forouzanfar et al. ⁴⁰ , WCRF and AICR ⁷⁸
Fiber	Colon/rectum ^a	0.809 (0.741–0.882) per 20 g day ⁻¹	Forouzanfar et al. ⁴⁰ , WCRF and AICR ⁷⁸
Fruit	Esophagus ^b	0.867 (0.776–0.968) per 100 g day ⁻¹	Forouzanfar et al. ⁴⁰ , Liu et al. ⁷⁹
Fruit	Trachea/bronchus/lung ^c	0.929 (0.890–0.970) per 100 g day ⁻¹	Forouzanfar et al. ⁴⁰ , Vieira et al. ⁸⁰
Milk	Colon/rectum ^a	0.898 (0.831–0.963) per 226.8 g day ⁻¹	Forouzanfar et al. ⁴⁰ , WCRF and AICR ⁷⁸
Processed meat	Colon/rectum ^a	1.179 (1.092–1.267) per 50 g day ⁻¹	Forouzanfar et al. ⁴⁰ , WCRF and AICR ⁷⁸
Red meat	Colon/rectum ^a	1.167 (1.033–1.309) per 100 g day ⁻¹	Forouzanfar et al. ⁴⁰ , WCRF and AICR ⁷⁸
Sodium	Stomach ^d	1.18 (1.02–1.38) per 1 g day ⁻¹ (e)	Forouzanfar et al. ⁴⁰ , WCRF and AICR ⁷⁸
Vegetables	Esophagus ^b	0.840 (0.780–0.920) per 100 g day ⁻¹ (f)	Forouzanfar et al. ⁴⁰ , Liu et al. ⁷⁹
Vitamin D	General	0.69 (0.55–0.86) per 1,429 IU day ⁻¹	Garland et al. ⁸¹

269 ^a 9.1% of total cancer mortality.⁸²
 270 ^b 2.4% of total cancer mortality.⁸²
 271 ^c 26.6% of total cancer mortality.⁸²
 272 ^d 1.9% of total cancer mortality.⁸²
 273

274 3. Results and Discussion

275 3.1 *Methylmercury exposures in 2014 among Lake Melville Inuit*

276 Measured Hg concentrations in the hair of Inuit individuals participating in our survey
277 ranged from 6.8 ng g⁻¹ to 6,200 ng g⁻¹. We find that between 67% (spring) and 71% (summer) of
278 all measured hair Hg samples fall within the modeled ranges of exposure (Figure 1, green
279 circles). This is better than many recent surveys^{47, 83-85} and may reflect lower inter-individual
280 variability in pharmacokinetics among a relatively homogeneous survey population and the
281 relatively smaller range of available fish with more consistent bioaccessibility.^{12, 86} The Inuit
282 Health Study (IHS) previously characterized MeHg exposures among Inuit in the community of
283 Rigolet and other communities on the Labrador Coast but excluding Happy Valley – Goose Bay
284 and North West River. The geometric mean blood Hg (3.2 µg L⁻¹) reported in the IHS is
285 equivalent to approximately 0.8 µg g⁻¹ hair^{87, 88} and compares well to Spring and Summer mean
286 hair levels (0.77 µg g⁻¹) measured in this study in Rigolet. MeHg exposures measured in 2014
287 are generally lower than other Inuit populations. For instance, the IHS reported geometric mean
288 blood Hg equal to 9.0 µg L⁻¹ for Inuit in Nunavut⁸⁹, while Dewailly et al.⁹⁰ reported a geometric
289 mean of 10.8 µg L⁻¹ for the Inuit of Nunavik (northern Quebec).

290 Exposures to MeHg of during the Summer survey period in 2014 were approximately
291 double the median exposure of the general Canadian population.^{2, 48} However, the majority of
292 individuals in the survey population fall below the reference dose (RfD) for exposure established
293 by the U.S. EPA of 0.1 µg kg⁻¹ day⁻¹ and Health Canada's provisional tolerable daily intake
294 (pTDI) of 0.2-0.47 µg kg⁻¹ day⁻¹ (Figure 2, SI Table S8).⁹¹ Across the three Inuit communities,
295 only individuals above the 90th percentile of MeHg exposures had daily intake levels that
296 exceeded the U.S. EPA's RfD (Figure 2). For example, the 95th percentile of MeHg exposures
297 ranged from 0.10 µg kg⁻¹ day⁻¹ in Happy Valley–Goose Bay to 0.27 µg kg⁻¹ day⁻¹ in Rigolet, and

298 0.12 $\mu\text{g kg}^{-1} \text{ day}^{-1}$ when averaged across the three communities. In Summer 2014, the fractions
299 of Lake Melville Inuit exceeding the Health Canada pTDI and the US EPA RfD were
300 approximately 1% and 7% respectively (SI Table S8). For all survey periods, older individuals,
301 men, and individuals residing in the community of Rigolet have higher MeHg exposures (SI
302 Table S8). Therefore, in 2014, risks associated with MeHg exposures were generally low and
303 concentrated among individuals at higher exposure percentiles.

304 Individuals with the highest MeHg exposures in 2014 had the highest intake of traditional
305 foods. Per-capita, roughly 70% of all MeHg intake came from traditional foods. Among
306 individuals with MeHg exposures $\geq 90^{\text{th}}$ percentile, 90% of all MeHg intake came from
307 traditional foods. In Summer 2014, individuals in the lowest quartile of traditional food intake (\leq
308 6.86 g day^{-1}) received 24% of their MeHg exposure from traditional foods compared to 80%
309 among the highest quartile ($>41.1 \text{ g day}^{-1}$). Median MeHg exposure was 0.043 $\mu\text{g kg}^{-1} \text{ day}^{-1}$ for
310 individuals in the highest quartile of traditional food consumption compared to 0.003 $\mu\text{g kg}^{-1}$
311 day^{-1} among individuals in the lowest intake quartile. Mean MeHg per-capita exposures in the
312 Summer survey period (0.035 $\mu\text{g kg}^{-1} \text{ day}^{-1}$) were significantly different from the Spring period
313 (0.024 $\mu\text{g kg}^{-1} \text{ day}^{-1}$, $p < 0.001$, Wilcox rank-sum test) but not the Winter (0.046 $\mu\text{g kg}^{-1} \text{ day}^{-1}$, p
314 >0.05). This mirrors trends in traditional food intake. Mean traditional food consumption was
315 significantly lower in the Spring period (28.5 g day^{-1}) compared to the Summer period (36.52 g
316 day^{-1} , $p < 0.001$, Wilcox rank-sum test). Mean traditional food consumption in Winter was 38.7
317 g day^{-1} but was not statistically different from either Spring or Summer. Therefore, while
318 population-wide MeHg exposure risks are generally low, these risks are sensitive to the MeHg
319 content of local foods.

320 Although locally caught Atlantic salmon is relatively low in MeHg (Table S3), it was the
321 single greatest contributor to overall MeHg intakes in all three communities including among
322 individuals with MeHg exposures $\geq 90^{\text{th}}$ percentile (24–29% of overall per-capita intake across
323 communities in Summer 2014). Other foods contributing more than 5% to overall MeHg intakes
324 in Summer 2014 were brook trout, Atlantic cod, tern eggs, duck and seal muscle (locally caught)
325 and fresh cod, canned tuna and fresh tuna (store-bought). Individuals with MeHg exposures $\geq 90^{\text{th}}$
326 percentile had similar sources of MeHg intake as the study population as a whole. SI Figure S2
327 presents the breakdown of per-capita MeHg sources in Summer 2014 for all individuals in the
328 three communities studied and for individuals in all communities with MeHg exposures $\geq 90^{\text{th}}$
329 percentile.

330 *3.2 2014 diet composition*

331 Survey data from the summer of 2014 suggest 91% of the population consumes
332 traditional foods. However, these foods are only account for approximately 2% of mean caloric
333 intake and 11% for the 95th percentile consumer (Figure 3). Across the three communities,
334 consumption of traditional foods is highest in Rigolet (mean = 4% of total calories) and lowest in
335 Happy Valley–Goose Bay (mean = <1% of total calories) (Figure 3). Prior work has reported
336 similar findings for other Inuit communities, with higher rates of country food consumption in
337 more northern communities that have less access to market alternatives.⁸⁸

338 Using food subsidy data, we estimate that store-bought nutrient-dense foods account for
339 25% of total per-capita caloric intake. SI Figure S3 presents the composition of this nutrient-
340 dense store-bought food component of the diet. For store-bought seafood, dietary survey data
341 agree to within 12% of estimates based on the food subsidy program data, providing partial
342 validation of this method. The remaining fractions of all caloric consumption estimated from
343 individual body weight must come from store bought food such as nutrient-sparse snack foods

344 and sweetened beverages. We estimate that these other foods account for approximately 70% of
345 all calories consumed across demographic groups. These findings agree with prior research that
346 has reported Inuit populations consume traditional foods, fruits and vegetables only one third as
347 frequently as foods with low nutrient content.⁹²

348 Types of traditional foods consumed by individuals vary widely (Figure 4). On a per-
349 capita basis across the three communities, 90% of the calories from traditional foods are derived
350 from 24 foods. Some traditional foods such as berries and seal blubber contain negligible
351 quantities of MeHg (Figure 4). Foods with negligible Hg account for 40% of the total calories
352 from traditional foods for all individuals surveyed (Figure 5). On a per-capita basis, 11 of the 24
353 major traditional foods consumed contain negligible amounts of MeHg. Survey data indicate that
354 the greatest diversity in traditional food consumption occurs among individuals with the highest
355 MeHg exposures (Figure 2b). Among the most highly exposed individuals to MeHg (90th
356 percentile), traditional foods with negligible MeHg content account for only 28% of total calories
357 (nine foods). Among all individuals, including among individuals with MeHg exposures greater
358 than the 90th percentile, the most widely consumed traditional foods with negligible Hg were
359 berries, goose, partridge, moose, caribou, seal blubber and rabbit.

360 Consumption of traditional foods increases linearly with age, with each year of age
361 associated with a 0.8 g day⁻¹ increase in traditional food intake ($R^2 = 0.10$, $p < 0.001$).
362 Traditional foods supply a significantly higher fraction of dietary calories for men (mean =
363 1.9%) compared to women (mean = 1.4%, $p < 0.001$, SI Table S9). Per-capita caloric
364 significance is reported in the SI for locally caught traditional seafood (Table S3), other locally
365 caught traditional food (Table S4), store-bought seafood (Table S5) and other store-bought
366 nutritious foods (Table S6).

3.3 *Importance of traditional foods for intake of nutrients*

367
368 Despite their low contribution to total caloric intake, traditional foods are the
369 predominant source of several key nutrients (Figure 5). Traditional foods supply approximately
370 70% of the n-3 PUFAs eicosapentaenoic acid (EPA) and docosahexaenoic acid (DHA) across the
371 population (Figure 5). They are an important source of vitamins D (35%), B12 (19%), B6 (6%),
372 A (16%), B2, B3 and C and iron (7%), and zinc (5%). Baseline dietary analysis suggests that
373 intake on average across the population of iron, n-3 fatty acids, vitamins A, B2 and B12 and zinc
374 are currently below dietary reference values.^{93, 94} Other studies have similarly found traditional
375 foods are richer in these nutrients than the market-based components of indigenous diets.⁹⁵⁻⁹⁸

3.4 *Nutritional impact of traditional food substitution scenarios*

376
377 While 2014 MeHg exposures were generally low, these exposures are likely to increase
378 as a result of upstream hydroelectric development.⁴² Food consumption advisories are commonly
379 used to control these risks but have unpredictable effects.^{4, 10, 11} Here, we describe the nutritional
380 impacts of several hypothetical responses to food consumption advisories.

381 Modeled dietary transitions to market foods, following the scenarios outlined above,
382 generally exacerbate deficiencies in n-3 PUFA and vitamins B12, D, B2, A, iron and zinc intake
383 among indigenous Inuit (Figure 5b). We find a small net gain in vitamin B-2 intake in the
384 nutrient-dense foods replacement scenario. Modeled reductions as a fraction of daily
385 recommended intake range from 1-2% for vitamin A to 37% for n-3 PUFAs. Replacement of
386 traditional foods with an equivalent amount of locally caught Atlantic salmon has a mixed
387 impact on nutritional sufficiency. Under the Atlantic salmon replacement scenario, average
388 intake of n-3 PUFAs and vitamin D increases by 53% and 10% of recommended daily values
389 respectively and leads to modest declines in intake of vitamins A and B-12, iron and zinc that are
390 less 5% of daily values (Figure 5b).

391 *3.5 Screening-level analysis of risks and benefits*

392 In 2014, intake of traditional foods among Inuit women of childbearing age generally had
393 a small net positive impact on child neurodevelopment due to relatively low MeHg levels in
394 traditional foods and benefits from n-3 PUFA intake. The median IQ decrement attributable to
395 present-day MeHg exposures from traditional foods is 0.02 points (95th population percentile:
396 0.14 points). After accounting for benefits from n-3 PUFAs in traditional foods, median (5th–95th
397 population percentiles) net impact on IQ is a gain of 0.01 points (decrement of 0.014 to gain of
398 0.19) IQ points. Increased consumption of greater quantities of low MeHg traditional foods
399 would further increase this net benefit.

400 Increased MeHg concentrations in local foods may pose neurodevelopmental and
401 cardiovascular risks. As described above, increased exposures are likely to disproportionately
402 impact individuals who already have high MeHg exposures and exceed regulatory reference
403 doses. However, mitigating these risks with reduced consumption of local foods may also pose
404 risks. Here, we present the results of our analysis comparing risks from increased MeHg
405 exposures to risks from potential dietary transitions. We conducted a screening-level estimate of
406 the risks (Figure 6) associated with peak forecasted increases in MeHg concentrations in
407 traditional foods due to hydroelectric flooding by Calder et al.⁴². Even at peak MeHg
408 concentrations, assuming the same magnitudes and species consumed as reported in the 2014
409 dietary survey, the population median IQ decrement for women of childbearing age is relatively
410 low (0.06 points). However, among individuals with MeHg exposures greater than the 90th
411 percentile, median impacts are much larger (loss of 1.4 IQ points). These estimates must be
412 viewed as uncertain due to inter-individual differences in sensitivity to MeHg exposure and the
413 toxicokinetics of MeHg absorption in the human body.^{86, 99}

414 Substitution of traditional foods with store-bought alternatives represents a large
415 reduction (70% per-capita) of n-3 PUFA intake (Figure 6). We estimate that this may also lead
416 to small neurodevelopmental impacts for most individuals (median IQ decrement of 0.01 across
417 substitution scenarios). However, replacement of traditional foods with locally caught Atlantic
418 salmon results in estimated gains of 0.08 IQ points (population-wide median) and 1.2 points
419 (>90th percentile MeHg exposures). This reflects a large increase in n-3 PUFA intake associated
420 with additional Atlantic Salmon consumption that greatly outweighs minor increases in MeHg
421 exposure (SI Table S9).

422 We estimate that cardiovascular risks associated with peak MeHg exposures in
423 traditional foods are smaller than under any store-bought food replacement scenario (Figure 6).
424 Across the population, median risk of cardiovascular mortality associated with projected
425 increases in MeHg in traditional foods increases by 3% relative to present-day (RR = 1.03). For
426 individuals with MeHg exposures at the 90th percentile or greater, estimated risk of
427 cardiovascular mortality increases dramatically (RR = 1.5). While the magnitude of these
428 impacts is highly uncertain due to variability in susceptibility to MeHg exposure across
429 populations, this analysis provides a quantitative estimate for the potential difference in
430 magnitude of risks from MeHg in comparison to dietary changes. If traditional foods are
431 replaced by processed meat, median RR of cardiovascular mortality is 1.19 across the population
432 and 3.14 for individuals with MeHg exposures >90th percentile. Median RR of cardiovascular
433 mortality is 1.08 at the across the population when the dietary replacement is fruits and
434 vegetables and 1.73 for individuals with MeHg exposures >90th percentile. The replacement
435 scenarios for nutrient-dense market foods and junk foods fall within this envelope (Figure 6).

436 Replacement of high MeHg traditional foods by locally caught Atlantic salmon has the
437 opposite impact of market foods on RR of cardiovascular mortality. This replacement scenario
438 leads to greater net benefits for cardiovascular health than all store-bought alternative scenarios.
439 Median RR of cardiovascular mortality under this scenario is 0.88 across the population and 0.55
440 for individuals with MeHg exposures >90th percentile (Figure 6, SI Table S10).

441 We estimate that replacing traditional foods with store-bought foods under all scenarios
442 will increase the RR of cancer mortality. Greater than 95% of this effect for the nutrient-dense
443 foods scenario is attributable to reduced intake of vitamin D. Median RR of colorectal cancer is
444 1.01 due to reduced fiber intake (75% of individuals) and increased consumption of red and
445 processed meat. Gains in calcium intake for 95% of individuals and increased milk consumption
446 do not offset these risks. Increased sodium intake (97% of individuals) results in a small increase
447 in RR of gastric cancer (RR = 0.01 at the 95th percentile). Colorectal cancers account for more
448 than three times as many deaths as gastric cancers in Newfoundland and Labrador¹⁰⁰, and so the
449 increased risk of colorectal cancer is a relatively stronger driver of overall cancer risks.

450 Replacement of the representative basket of traditional foods with locally caught Atlantic salmon
451 provides small net reductions in overall cancer risks relative to present-day (median RR = 0.995;
452 >90th percentile individuals: RR = 0.97) (Figure 6, SI Table S11).

453 3.6 Study strengths and limitations

454 To our knowledge, this study is the most comprehensive survey of Inuit diet and MeHg
455 exposures ($n = 1,145$). We provide a detailed characterization of diet variability among Lake
456 Melville Inuit evaluated with direct hair Hg measurements. Our dietary MeHg exposure model
457 performed better than several other recent studies, likely reflecting the relatively homogeneous
458 sources of MeHg across the population and the use of extensive local data. This study provides
459 an assessment of the magnitude of projected risks associated with elevated MeHg exposures in

460 comparison with the neurodevelopmental, cardiovascular, cancer and nutritional risks posed by
461 possible dietary changes.

462 We were limited by the availability of intake data for store-bought foods (other than
463 seafood, which we measured on the FFQ), which are available only on a per-capita basis for the
464 community of Rigolet. Therefore, our characterization of present-day nutritional sufficiency and
465 composition of store-bought foods could not describe interindividual variability. Dietary recall
466 data is often biased, and although we designed the study so as to evaluate (with hair Hg
467 measurements) and control for some biases (e.g., correcting for species-specific recall biases by
468 asking redundant “total” recall questions) as described above, we are limited by the accuracy of
469 the reports of survey respondents. Our evaluation of dietary model performance with respect to
470 hair Hg and calculation of MeHg exposures from hair Hg measurements depends on self-
471 reported measures of height and weight, which may be estimates. Although there was little
472 evident bias in the larger-scale Summer survey round (Figure 1), these factors likely contributed
473 to the random error observed.

474 Our analysis does not account for other possible second-order effects of traditional food
475 substitutions. For instance, isocaloric dietary substitution is associated with a mean reduction in
476 protein intake of 2–11% across dietary scenarios. While per-capita intake of protein continues to
477 exceed the recommended daily allowance, increasing the proportion of calories from
478 carbohydrates and fats may lead to overall greater caloric intake via reduced satiety and higher
479 insulin production, thus increasing weight gain and obesity-related risks.^{101, 102} Consumption
480 rates of locally caught traditional foods in the Inuit communities studied here are lower than
481 those for other indigenous communities across Canada. For example, British Columbia First
482 Nations consume roughly three times the per-capita amounts reported here.¹⁰³ This implies

483 health and nutritional impacts associated with dietary transitions may be greater in other
484 populations.

485 We have not addressed the physical or psychosocial dimensions of hunting and fishing
486 and the preparation and consumption of traditional foods. Hunting traditional foods represents
487 vigorous physical activity, and loss of access to traditional foods has been linked to adverse
488 mental and social outcomes, implying substitution of traditional foods may present additional
489 risks to those quantified here.¹⁰⁴⁻¹⁰⁶ Fruits and vegetables are not a significant part of the
490 traditional Inuit diet¹⁰⁷, and our analysis suggests they account for roughly 2% of caloric intake
491 at present day. Replacement of traditional foods with fruits and vegetables is acknowledged to be
492 less likely than by other foods such as red meat or other snack foods. This scenario is included as
493 a better-case scenario for store-bought alternatives.

494 3.7 Implications for risk mitigation strategies

495 Food consumption advisories are routinely used to mitigate potential risks from elevated
496 contaminant exposures. However, these advisories have unpredictable effects and can lead to
497 reduced overall intake of traditional foods among indigenous populations. This study is the first
498 to calculate the plausible range of health impacts from elevated MeHg exposures as compared to
499 potential outcomes of risk-mitigation strategies. Our analysis suggests that replacing traditional
500 foods with store-bought alternatives may lead to increases in cardiovascular and cancer risks
501 among Lake Melville Inuit. Conversely, we estimate that replacement with locally caught
502 Atlantic salmon will lead to net benefits for neurodevelopmental and cardiovascular health and
503 reduce cancer risks relative to the present-day diet. Atlantic salmon is already a large component
504 of traditional diet of our survey respondents, accounting for approximately 25% of calories from
505 traditional foods. These results reinforce the potential benefits of dietary advice that promotes
506 nutrient-dense, low-MeHg traditional foods. We have shown that in the local diet, there are many

507 commonly consumed (and therefore familiar) foods with negligible levels of Hg, intake of which
508 could be promoted in order to maximize net health benefits of the traditional diet.

509 Reducing the diversity of traditional foods consumed has mixed impacts on nutritional
510 sufficiency, which must be considered when making recommendations about dietary choices
511 among indigenous populations. Nutrient shortfalls are common in indigenous populations, and
512 our findings suggest that this is the case among Lake Melville Inuit. Therefore, independent of
513 contaminant levels, there may be a role for dietary interventions that promote increased intake of
514 nutritious foods and possibly dietary supplements. Taken together, findings presented here
515 underline the importance of protecting northern food webs from environmental contamination
516 and of promoting traditional foods among indigenous populations.

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References

1. Arctic Monitoring and Assessment Program (AMAP), *AMAP Assessment 2015: Human Health in the Arctic*. Oslo, Norway, **2015**.
2. Lye, E.; Legrand, M.; Clarke, J.; Probert, A., Blood total mercury concentrations in the Canadian population: Canadian health measures survey cycle 1, 2007-2009. *Can J Public Health* **2013**, *104*, (3), e246-e251. DOI: <https://www.jstor.org/stable/canajpublhealth.104.3.e246>
3. Van Oostdam, J.; Donaldson, S. G.; Feeley, M.; Arnold, D.; Ayotte, P.; Bondy, G., . . . Kalhok, S., Human health implications of environmental contaminants in Arctic Canada: A review. *Sci Total Environ* **2005**, *351-352*, 165-246. DOI: 10.1016/j.scitotenv.2005.03.034
4. Passos, C. J.; Mergler, D., Human mercury exposure and adverse health effects in the Amazon: a review. *Cad Saude Publica* **2008**, *24*, S503-S520. DOI: 10.1590/S0102-311X2008001600004
5. Hydro-Québec; Conseil cri de la santé et des services sociaux de la Baie James, *Le guide alimentaire des poissons nordiques. Région de la Baie-James*. **2013**.
6. Hydro-Québec Production, *Guide de consommation des poissons pour les plans d'eau de la région de la rivière Saint-Maurice en Haute-Mauricie*. **2014**.
7. Agence de la santé et des services sociaux de la Côte-Nord; Hydro-Québec Production; CHU de Québec; Institute national de santé publique, *Le guide alimentaire des poissons et fruits de mer de la Côte-Nord*. **2013**.
8. Shimshack, J. P.; Ward, M. B., Mercury advisories and household health trade-offs. *J Health Econ* **2010**, *29*, (5), 674-85. DOI: 10.1016/j.jhealeco.2010.05.001
9. Teisl, M. F.; Fromberg, E.; Smith, A. E.; Boyle, K. J.; Engelberth, H. M., Awake at the switch: improving fish consumption advisories for at-risk women. *Sci Total Environ* **2011**, *409*, (18), 3257-66. DOI: 10.1016/j.scitotenv.2011.05.006
10. Furgal, C.; Powell, S.; Myers, H., Digesting the message about contaminants and country foods in the Canadian North: a review and recommendations for future research and action. *Arctic* **2005**, *58*, (2), 103-114. DOI: 10.14430/arctic404
11. Wheatley, B.; Paradis, S., Balancing human exposure, risk and reality: questions raised by the Canadian aboriginal methylmercury program. *Neurotoxicology* **1996**, *17*, (1), 241-249.
12. Li, M.; von Stackelberg, K.; Rheinberger, C. M.; Hammitt, J. K.; Krabbenhoft, D. P.; Yin, R.; Sunderland, E. M., Insights from mercury stable isotopes into factors affecting the internal body burden of methylmercury in frequent fish consumers. *Elementa* **2016**, *4*, 000103. DOI: 10.12952/journal.elementa.000103
13. Binnington, M. J.; Quinn, C. L.; McLachlan, M. S.; Wania, F., Evaluating the effectiveness of fish consumption advisories: modeling prenatal, postnatal, and childhood

exposures to persistent organic pollutants. *Environ Health Perspect* **2014**, *122*, (2), 178-86. DOI: 10.1289/ehp.1206380

14. Karagas, M.; Choi, A. L.; Oken, E.; Horvat, M.; Schoeny, R.; Kamai, E., . . . Korrick, S., Evidence on the human health effects of low level methylmercury exposure. *Environ Health Perspect* **2012**, *120*, (6), 799-806. DOI: 10.1289/ehp.1104494

15. Debes, F.; Weihe, P.; Grandjean, P., Cognitive deficits at age 22 years associated with prenatal exposure to methylmercury. *Cortex* **2016**, *74*, 358-69. DOI: 10.1016/j.cortex.2015.05.017

16. United States Environmental Protection Agency (US EPA), *Methylmercury (MeHg); CASRN 22967-92-6*. Integrated Risk Information System (IRIS). Washington, DC, **2002**.

17. Boucher, O.; Jacobson, S.; Plusquellec, P.; Dewailly, E.; Ayotte, P.; Forget-Dubois, N., . . . Muckle, G., Prenatal methylmercury, postnatal lead exposure, and evidence of attention deficit/hyperactivity disorder among Inuit children in Arctic Quebec. *Environ Health Perspect* **2012**, *120*, (10), 1456-1461. DOI: 10.1289/ehp.1204976

18. Weihe, P.; Hansen, J. C.; Katsuyuki, M.; Debes, F.; Jorgensen, P. J.; Steuerwald, U., . . . Grandjean, P., Neurobehavioral performance of Inuit children with increased prenatal exposure to methylmercury. *Int J Circumpol Heal* **2016**, *61*, 41-49. DOI: 10.3402/ijch.v61i0.17404

19. Virtanen, J. K.; Rissanen, T. H.; Voutilainen, S.; Tuomainen, T. P., Mercury as a risk factor for cardiovascular diseases. *J Nutr Biochem* **2007**, *18*, (2), 75-85. DOI: 10.1016/j.jnutbio.2006.05.001

20. Farina, M.; Rocha, J. B.; Aschner, M., Mechanisms of methylmercury-induced neurotoxicity: evidence from experimental studies. *Life Sci* **2011**, *89*, (15-16), 555-63. DOI: 10.1016/j.lfs.2011.05.019

21. Salonen, J. T.; Seppanen, K.; Nyssonen, K.; Korpela, H.; Kauhanen, J.; Kantola, M., . . . Salonen, R., Intake of mercury from fish, lipid peroxidation, and the risk of myocardial infarction and coronary, cardiovascular, and any death in Eastern Finnish men. *Circulation* **1995**, *91*, (3), 645-655. DOI: 10.1161/01.cir.91.3.645

22. Roman, H. A.; Walsh, T. L.; Coull, B. A.; Dewailly, E.; Guallar, E.; Hattis, D., . . . Rice, G., Evaluation of the cardiovascular effects of methylmercury exposures: current evidence supports development of a dose-response function for regulatory benefits analysis. *Environ Health Perspect* **2011**, *119*, (5), 607-14. DOI: 10.1289/ehp.1003012

23. Laird, B. D.; Goncharov, A. B.; Egeland, G. M.; Chan, H. M., Dietary advice on Inuit traditional food use needs to balance benefits and risks of mercury, selenium, and n3 fatty acids. *J Nutr* **2013**, *143*, (6), 923-30. DOI: 10.3945/jn.112.173351

24. European Food Safety Authority (EFSA), Scientific opinion on health benefits of seafood (fish and shellfish) consumption in relation to health risks associated with exposure to methylmercury. *EFSA Journal* **2014**, *12*, (7). DOI: 10.2903/j.efsa.2014.3761

25. Ginsberg, G. L.; Toal, B. F., Quantitative approach for incorporating methylmercury risks and omega-3 fatty acid benefits in developing species-specific fish consumption advice. *Environ Health Perspect* **2009**, *117*, (2), 267-75. DOI: 10.1289/ehp.11368
26. Stern, A. H.; Korn, L. R., An approach for quantitatively balancing methylmercury risk and omega-3 benefit in fish consumption advisories. *Environ Health Perspect* **2011**, *119*, (8), 1043-1046. DOI: 10.1289/ehp1002824
27. Mahaffey, K. R.; Sunderland, E. M.; Chan, H. M.; Choi, A. L.; Grandjean, P.; Mariën, K., . . . Yasutake, A., Balancing the benefits of n-3 polyunsaturated fatty acids and the risks of methylmercury exposure from fish consumption. *Nutr Rev* **2011**, *69*, (9), 493-508. DOI: 10.1111/j.1753-4887.2011.00415.x
28. Food and Agricultural Organization of the United Nations (FAO); World Health Organization (WHO), *Joint FAO/WHO Expert Consultation on the Risks and Benefits of Fish Consumption, 25-29 January 2010, Rome, Italy*. World Health Organization: Geneva, Switzerland, **2010**.
29. Threapleton, D. E.; Greenwood, D. C.; Evans, C. E.; Cleghorn, C. L.; Nykjaer, C.; Woodhead, C., . . . Burley, V. J., Dietary fibre intake and risk of cardiovascular disease: systematic review and meta-analysis. *Brit Med J* **2013**, *347*, f6879. DOI: 10.1136/bmj.f6879
30. Whelton, P. K.; Appel, L. J.; Sacco, R. L.; Anderson, C. A.; Antman, E. M.; Campbell, N., . . . Van Horn, L. V., Sodium, blood pressure, and cardiovascular disease: further evidence supporting the American Heart Association sodium reduction recommendations. *Circulation* **2012**, *126*, (24), 2880-9. DOI: 10.1161/CIR.0b013e318279acbf
31. Koeth, R. A.; Wang, Z.; Levison, B. S.; Buffa, J. A.; Org, E.; Sheehy, B. T., . . . Hazen, S. L., Intestinal microbiota metabolism of L-carnitine, a nutrient in red meat, promotes atherosclerosis. *Nat Med* **2013**, *19*, (5), 576-85. DOI: 10.1038/nm.3145
32. Wang, X.; Ouyang, Y.; Liu, J.; Zhu, M.; Zhao, G.; Bao, W.; Hu, F. B., Fruit and vegetable consumption and mortality from all causes, cardiovascular disease, and cancer: systematic review and dose-response meta-analysis of prospective cohort studies. *Brit Med J* **2014**, *349*, g4490. DOI: 10.1136/bmj.g4490
33. Blomhoff, R.; Carlsen, M. H.; Andersen, L. F.; Jacobs, D. R., Health benefits of nuts: potential role of antioxidants. *Brit J Nutr* **2006**, *96*, (S2), S52-S60. DOI: 10.1017/BJN20061864
34. Ludwig, D. S.; Pereira, M. A.; Kroenke, C. H.; Hilner, J. E.; Van Horn, L.; Slattery, M. L.; Jacobs Jr, D. R., Dietary fiber, weight gain, and cardiovascular disease risk factors in young adults. *J Amer Med Assoc* **1999**, *282*, (16), 1539-1546. DOI: 10.1001/jama.282.16.1539
35. Kelly, J. H.; Sabaté, J., Nuts and coronary heart disease: an epidemiological perspective. *Brit J Nutr* **2006**, *96*, (S2), S61-S67. DOI: 10.1017/BJN20061865
36. Rissanen, T. H.; Voutilainen, S.; Virtanen, J. K.; Venho, B.; Vanharanta, M.; Mursu, J.; Salonen, J. T., Low intake of fruits, berries and vegetables is associated with excess mortality in

- men: the Kuopio Ischaemic Heart Disease Risk Factor (KIHD) Study. *J Nutr* **2003**, *133*, (1), 199-204. DOI: 10.1093/jn/133.1.199
37. Giovannucci, E., The epidemiology of vitamin D and cancer incidence and mortality: a review (United States). *Cancer Cause Control* **2005**, *16*, (2), 83-95. DOI: 10.1007/s10552-004-1661-4
38. Giovannucci, E.; Liu, Y.; Rimm, E. B.; Hollis, B. W.; Fuchs, C. S.; Stampfer, M. J.; Willett, W. C., Prospective study of predictors of vitamin D status and cancer incidence and mortality in men. *J Natl Cancer I* **2006**, *98*, (7), 451-459. DOI: 10.1093/jnci/djj101
39. Grant, W. B., Relation between prediagnostic serum 25-hydroxyvitamin D level and incidence of breast, colorectal, and other cancers. *J Photoch Photobio B* **2010**, *101*, (2), 130-6. DOI: 10.1016/j.jphotobiol.2010.04.008
40. Forouzanfar, M. H.; Afshin, A.; Alexander, L. T.; Anderson, H. R.; Bhutta, Z. A.; Biryukov, S., . . . Murray, C. J. L., Global, regional, and national comparative risk assessment of 79 behavioural, environmental and occupational, and metabolic risks or clusters of risks, 1990–2015: a systematic analysis for the Global Burden of Disease Study 2015. *Lancet* **2016**, *388*, (10053), 1659-1724. DOI: 10.1016/S0140-6736(16)31679-8
41. Wang, H.; Naghavi, M.; Allen, C.; Barber, R. M.; Bhutta, Z. A.; Carter, A., . . . Coates, M. M., Global, regional, and national life expectancy, all-cause mortality, and cause-specific mortality for 249 causes of death, 1980–2015: a systematic analysis for the Global Burden of Disease Study 2015. *Lancet* **2016**, *388*, (10053), 1459-1544. DOI: 10.1016/S0140-6736(16)31012-1
42. Calder, R. S. D.; Schartup, A. T.; Li, M.; Valberg, A. P.; Balcom, P. H.; Sunderland, E. M., Future impacts of hydroelectric power development on methylmercury exposures of Canadian indigenous communities. *Environ Sci Technol* **2016**, *50*, (23), 13115-13122. DOI: 10.1021/acs.est.6b04447
43. Statistics Canada, Census Profile. 2011 Census. Accessed 2018-03-18 from <http://www12.statcan.gc.ca/census-recensement/2011/dp-pd/prof/index.cfm?Lang=E>.
44. Durkalec, A.; Sheldon, T.; Bell, T., Eds., Scientific Report. Lake Melville: Avativut, Kanuittailinnivut. Nunatsiavut Government: Nain, NL, **2016**.
45. Durkalec, A.; Sheldon, T.; Bell, T., Eds., Summary for Policymakers. Lake Melville: Avativut, Kanuittailinnivut. Nunatsiavut Government: Nain, NL, **2016**.
46. Li, M.; Schartup, A. T.; Valberg, A. P.; Ewald, J. D.; Krabbenhoft, D. P.; Yin, R., . . . Sunderland, E. M., Environmental origins of methylmercury accumulated in subarctic estuarine fish indicated by mercury stable isotopes. *Environ Sci Technol* **2016**, *50*, (21), 11559-11568. DOI: 10.1021/acs.est.6b03206

47. Lincoln, R. A.; Shine, J. P.; Chesney, E. J.; Vorhees, D. J.; Grandjean, P.; Senn, D. B., Fish consumption and mercury exposure among Louisiana recreational anglers. *Environ Health Perspect* **2011**, *119*, (2), 245-51. DOI: 10.1289/ehp.1002609
48. Stern, A. H., Estimation of the interindividual variability in the one-compartment pharmacokinetic model for methylmercury: implications for the derivation of a reference dose. *Regul Toxicol Pharm* **1997**, *25*, (3), 277-288. DOI: 10.1006/rtph.1997.1105
49. Aboriginal Affairs and Northern Development Canada (AANDC), *Access to Information Act: request A-2016-00810/VN*. **2016**.
50. Douglass, J. S.; Fleming, K. H.; Barraji, L. M.; Heimbach, J. T., Using Food Consumption Data to Determine Exposure to Toxins in *Handbook of Human Toxicology*, Massaro, E. J., Ed. CRC Press: Boca Raton, FL, 1997; pp 305-326.
51. Sunderland, E. M., Mercury exposure from domestic and imported estuarine and marine fish in the U.S. seafood market. *Environ Health Perspect* **2007**, *115*, (2), 235-42. DOI: 10.1289/ehp.9377
52. Sunderland, E. M.; Li, M.; Bullard, K., Decadal changes in the edible supply of seafood and methylmercury exposure in the United States. *Environ Health Perspect* **2018**, *126*, (1), 017006. DOI: 10.1289/EHP2644
53. Government of Canada, Nutrition North. Accessed 2017-06-23 from www.nutritionnorthcanada.gc.ca/eng/1415385762263/1415385790537.
54. Gustavsson, J.; Cederberg, C.; Sonersson, U.; van Otterdijk, R.; Meybeck, A., *Global Food Losses and Food Waste: Extent, Causes and Prevention*. United Nations Food and Agriculture Organization (UN FAO): Rome, Italy, **2011**.
55. Mifflin, M. D.; St Jeor, S. T.; Hill, L. A.; Scott, B. J.; Daugherty, S. A.; Koh, Y., A new predictive equation for resting energy expenditure in healthy individuals. *Am J Clin Nutr* **1990**, *51*, (2), 241-247. DOI: 10.1093/ajcn/51.2.241
56. Kattelman, K. K.; Conti, K.; Ren, C., The Medicine Wheel nutrition intervention: a diabetes education study with the Cheyenne River Sioux Tribe. *J Am Diet Assoc* **2010**, *110*, (5 Suppl), S44-51. DOI: 10.1016/j.jada.2010.03.003
57. Sheehy, T.; Kolahdooz, F.; Roache, C.; Sharma, S., Changing dietary patterns in the Canadian Arctic: Frequency of consumption of foods and beverages by. *Food Nutr Bull* **2014**, *35*, (2), 244-252. DOI: 10.1177/156482651403500211
58. Health Canada, Canadian Nutrient File. Accessed 2018-01-15 from <https://food-nutrition.canada.ca/cnf-fce/index-eng.jsp>.
59. Rochette, L.; Blanchet, C., *Methodological Report. Qanuippitaa? (How Are We?)*. Institut national de santé publique du Québec: Québec, QC, **2004**.

60. United States Department of Agriculture (USDA), Food Composition Databases. Accessed 2017-06-16 from <https://ndb.nal.usda.gov/ndb/>.
61. Kuhnlein, H. V.; Receveur, O.; Soueida, R.; Berti, P. R., Unique patterns of dietary adequacy in three cultures of Canadian Arctic indigenous peoples. *Public Health Nutr* **2008**, *11*, (4), 349-60. DOI: 10.1017/S1368980007000353
62. Schaefer, S. E.; Erber, E.; Trzaskos, J. P.; Roache, C.; Osborne, G.; Sharma, S., Sources of Food Affect Dietary Adequacy of Inuit Women of Childbearing Age in Arctic Canada. *J Health Popul Nutr* **2011**, *29*, (5). DOI: 10.3329/jhpn.v29i5.8899
63. Lepage, C.; Carignan, G.; Patry, P.; Saucier, A.; Jetté, M., Use of Services and Consumption of Medications in *A Health Profile of the Inuit*, Jetté, M., Ed. Santé Québec: Montréal, QC.
64. McDowell, M. A.; Dillon, C. F.; Osterloh, J.; Bolger, P. M.; Pellizzari, E.; Fernando, R., . . . Mahaffey, K. R., Hair Mercury Levels in U.S. Children and Women of Childbearing Age: Reference Range Data from NHANES 1999-2000. *Environ Health Perspect* **2004**, *112*, (11), 1165-1171. DOI: 10.1289/ehp.7046
65. Fleming, S. T.; Rastogi, A.; Dmitrienko, A.; Johnson, K. D., A comprehensive prognostic index to predict survival based on multiple comorbidities: a focus on breast cancer. *Med Care* **1999**, 601-614. DOI: <https://www.jstor.org/stable/3767021>
66. Forouzanfar, M. H.; Alexander, L.; Anderson, H. R.; Bachman, V. F.; Biryukov, S.; Brauer, M., . . . Murray, C. J., Global, regional, and national comparative risk assessment of 79 behavioural, environmental and occupational, and metabolic risks or clusters of risks in 188 countries, 1990-2013: a systematic analysis for the Global Burden of Disease Study 2013. *Lancet* **2015**, *386*, (10010), 2287-2323. DOI: 10.1016/S0140-6736(15)00128-2
67. Gates, A.; Skinner, K.; Gates, M., The diets of school-aged Aboriginal youths in Canada: a systematic review of the literature. *J Hum Nutr Diet* **2015**, *28*, (3), 246-61. DOI: 10.1111/jhn.12246
68. Kuhnlein, H. V.; Chan, H. M., Environment and contaminants in traditional food systems of northern indigenous peoples. *Ann Rev Nutr* **2000**, *20*, (1), 595-626. DOI: 10.1146/annurev.nutr.20.1.595
69. Grant, W. B.; Schwalfenberg, G. K.; Genuis, S. J.; Whiting, S. J., An estimate of the economic burden and premature deaths due to vitamin D deficiency in Canada. *Mol Nutr Food Res* **2010**, *54*, (8), 1172-81. DOI: 10.1002/mnfr.200900420
70. Cohen, J. T.; Bellinger, D. C.; Connor, W. E.; Shaywitz, B. A., A quantitative analysis of prenatal intake of n-3 polyunsaturated fatty acids and cognitive development. *Am J Prev Med* **2005**, *29*, (4), 366-74. DOI: 10.1016/j.amepre.2005.06.008
71. Chowdhury, R.; Stevens, S.; Gorman, D.; Pan, A.; Warnakula, S.; Chowdhury, S., . . . Hu, F. B., Association between fish consumption, long chain omega 3 fatty acids, and risk of

cerebrovascular disease: systematic review and meta-analysis. *Brit Med J* **2012**, *345*, e6698. DOI: 10.1136/bmj.e6698

72. Hu, D.; Huang, J.; Wang, Y.; Zhang, D.; Qu, Y., Fruits and vegetables consumption and risk of stroke: a meta-analysis of prospective cohort studies. *Stroke* **2014**, *45*, (6), 1613-9. DOI: 10.1161/STROKEAHA.114.004836

73. Afshin, A.; Micha, R.; Khatibzadeh, S.; Mozaffarian, D., Consumption of nuts and legumes and risk of incident ischemic heart disease, stroke, and diabetes: a systematic review and meta-analysis. *Am J Clin Nutr* **2014**, *100*, (1), 278-288. DOI: 10.3945/ajcn.113.076901

74. Micha, R.; Wallace, S. K.; Mozaffarian, D., Red and processed meat consumption and risk of incident coronary heart disease, stroke, and diabetes mellitus: a systematic review and meta-analysis. *Circulation* **2010**, *121*, (21), 2271-83. DOI: 10.1161/CIRCULATIONAHA.109.924977

75. Mozaffarian, D.; Clarke, R., Quantitative effects on cardiovascular risk factors and coronary heart disease risk of replacing partially hydrogenated vegetable oils with other fats and oils. *Eur J Clin Nutr* **2009**, *63*, (S2), S22. DOI: 10.1038/sj.ejcn.1602976

76. Centers for Disease Control and Prevention (CDC), *Death rates for 358 selected causes, by 10-year age groups, race, and sex: United States, 1999-2007*. National Vital Statistics System. **2010**.

77. Drewnowski, A., Sensory properties of fats and fat replacements. *Nutr Rev* **1992**, *50*, (4), 17-20. DOI: 10.1111/j.1753-4887.1992.tb01285.x

78. World Cancer Research Fund (WCRF); American Institute for Cancer Research (AICR), *Food, Nutrition, Physical Activity, and the Prevention of Cancer: a Global Perspective*. AICR: Washington, DC, **2007**.

79. Liu, J.; Wang, J.; Leng, Y.; Lv, C., Intake of fruit and vegetables and risk of esophageal squamous cell carcinoma: a meta-analysis of observational studies. *Int J Cancer* **2013**, *133*, (2), 473-85. DOI: 10.1002/ijc.28024

80. Vieira, A. R.; Abar, L.; Vingeliene, S.; Chan, D.; Aune, D.; Navarro-Rosenblatt, D., . . . Norat, T., Fruits, vegetables and lung cancer risk: a systematic review and meta-analysis. *Ann Oncol* **2015**, *27*, (1), 81-96. DOI: 10.1093/annonc/mdv381

81. Garland, C. F.; Grant, W. B.; Mohr, S. B.; Gorham, E. D.; Garland, F. C., What is the Dose-Response Relationship between Vitamin D and Cancer Risk? *Nutr Rev* **2007**, *65*, (8), 91-95. DOI: 10.1301/nr.2007.aug.S91-S95

82. Howlander, N.; Noone, A.; Krapcho, M.; Miller, D.; Bishop, K.; Altekruse, S., . . . Cronin, K., Eds., *SEER Cancer Statistics Review, 1975-2013*. Bethesda, MD, **2016**.

83. Canuel, R.; de Grosbois, S. B.; Atikessé, L.; Lucotte, M.; Arp, P.; Ritchie, C., . . . Anderson, R., New evidence on variations of human body burden of methylmercury from fish consumption. *Environ Health Perspect* **2006**, *114*, (2), 302-306. DOI: 10.1289/ehp.7857
84. Gosselin, N. H.; Brunet, R. C.; Carrier, G.; Bouchard, M.; Feeley, M., Reconstruction of methylmercury intakes in indigenous populations from biomarker data. *J Expo Sci Env Epid* **2006**, *16*, (1), 19-29. DOI: 10.1038/sj.jea.7500433
85. Sirot, V.; Guerin, T.; Mauras, Y.; Garraud, H.; Volatier, J. L.; Leblanc, J. C., Methylmercury exposure assessment using dietary and biomarker data among frequent seafood consumers in France CALIPSO study. *Environ Res* **2008**, *107*, (1), 30-8. DOI: 10.1016/j.envres.2007.12.005
86. Basu, N.; Goodrich, J. M.; Head, J., Ecogenetics of mercury: from genetic polymorphisms and epigenetics to risk assessment and decision-making. *Environ Toxicol Chem* **2014**, *33*, (6), 1248-58. DOI: 10.1002/etc.2375
87. World Health Organization (WHO), *Methylmercury*. Environmental Health Criteria. Geneva, Switzerland, **1990**.
88. Chan, H. M., *Contaminant Assessment in Nunatsiavut*. Inuit Health Survey 2007-2008. **2011**.
89. Chan, H. M., *Contaminant Assessment in Nunavut*. Inuit Health Survey 2007-2008. **2011**.
90. Dewailly, E.; Ayotte, P.; Pereg, D.; Dery, S.; Dallaire, R.; Fontaine, J.; Côté, S., Exposure to environmental contaminants in Nunavik: metals. *Nunavik Inuit Health Survey 2004* **2004**.
91. Health Canada, Mercury: Your Health and the Environment: A Resource Tool. Accessed 2017-11-21 from <https://www.canada.ca/en/health-canada/services/environmental-workplace-health/reports-publications/environmental-contaminants/mercury-your-health-environment-resource-tool.html>.
92. Hopping, B. N.; Erber, E.; Mead, E.; Sheehy, T.; Roache, C.; Sharma, S., Socioeconomic indicators and frequency of traditional food, junk food, and fruit and vegetable consumption amongst Inuit adults in the Canadian Arctic. *J Hum Nutr Diet* **2010**, *23 Suppl 1*, 51-8. DOI: 10.1111/j.1365-277X.2010.01100.x
93. Kris-Etherton, P. M.; Grieger, J. A.; Etherton, T. D., Dietary reference intakes for DHA and EPA. *Prostag Leukotr Ess* **2009**, *81*, (2), 99-104. DOI: 10.1016/j.plefa.2009.05.011
94. United States Food and Drug Administration (US FDA), *A Food Labeling Guide: Guidance for Industry*. College Park, MD, **2013**.
95. Gagné, D.; Blanchet, R.; Lauzière, J.; Vaissière, E.; Vézina, C.; Ayotte, P., . . . Turgeon O'Brien, H., Traditional food consumption is associated with higher nutrient intakes in Inuit

children attending childcare centres in Nunavik. *Int J Circumpol Heal* **2012**, *71*, 18401. DOI: 10.3402/ijch.v71i0.18401

96. Kuhnlein, H.; Receveur, O., Local cultural animal food contributes high levels of nutrients for arctic Canadian indigenous adults and children. *J Nutr* **2007**, *137*, (4), 1110-1114. DOI: 10.1093/jn/137.4.1110

97. Nakano, T.; Fediuk, K.; Kassi, N.; Kuhnlein, H. V., Food use of Dene/Metis and Yukon children. *Int J Circumpol Heal* **2005**, *64*, (2). DOI: 10.3402/ijch.v64i2.17966

98. Sheehy, T.; Kolahdooz, F.; Roache, C.; Sharma, S., Traditional food consumption is associated with better diet quality and adequacy among Inuit adults in Nunavut, Canada. *Int J Food Sci Nutr* **2015**, *66*, (4), 445-51. DOI: 10.3109/09637486.2015.1035232

99. Grandjean, P.; Budtz-Jorgensen, E., An ignored risk factor in toxicology: The total imprecision of exposure assessment. *Pure Appl Chem* **2010**, *82*, (2), 383-391. DOI: 10.1351/PAC-CON-09-05-04

100. Statistics Canada, Age-standardized five-year survival estimates for primary sites of cancer, ICD-O-3 (October 2011 CCR file), by sex, 3 years of cases, Canada and selected provinces. CANSIM. Accessed 2018-03-19.

101. Simpson, S. J.; Raubenheimer, D., *The Nature of Nutrition: A Unifying Framework from Animal Adaptation to Human Obesity*. Princeton University Press: Princeton, NJ, **2012**.

102. Mozaffarian, D.; Hao, T.; Rimm, E. B.; Willett, W. C.; Hu, F. B., Changes in diet and lifestyle and long-term weight gain in women and men. *New Engl J Med* **2011**, *364*, (25), 2392-404. DOI: 10.1056/NEJMoa1014296

103. Chan, H. M.; Receveur, O.; Sharp, D.; Schwartz, H.; Ing, A.; Tikhonov, C., *First Nations Food, Nutrition and Environment Study: Results from British Columbia (2008/2009)*. University of Northern British Columbia: Prince George, BC, **2011**.

104. Kirmayer, L. J.; Fletcher, C.; Watt, R., Locating the Ecocentric Self: Inuit Concepts of Mental Health and Illness in *Healing Traditions: The Mental Health of Aboriginal Peoples in Canada*, Kirmayer, L. J., Ed. UBC Press: Vancouver, BC, 2009.

105. King, M.; Smith, A.; Gracey, M., Indigenous health part 2: the underlying causes of the health gap. *Lancet* **2009**, *374*, (9683), 76-85. DOI: 10.1016/S0140-6736(09)60827-8

106. Sharma, S., Assessing diet and lifestyle in the Canadian Arctic Inuit and Inuvialuit to inform a nutrition and physical activity intervention programme. *J Hum Nutr Diet* **2010**, *23 Suppl 1*, 5-17. DOI: 10.1111/j.1365-277X.2010.01093.x

107. Cordain, L.; Eaton, S. B.; Miller, J. B.; Mann, N.; Hill, K., The paradoxical nature of hunter-gatherer diets: meat-based, yet non-atherogenic. *Eur J Clin Nutr* **2002**, *56*, (S1), S42. DOI: 10.1038/sj.ejcn.1601353

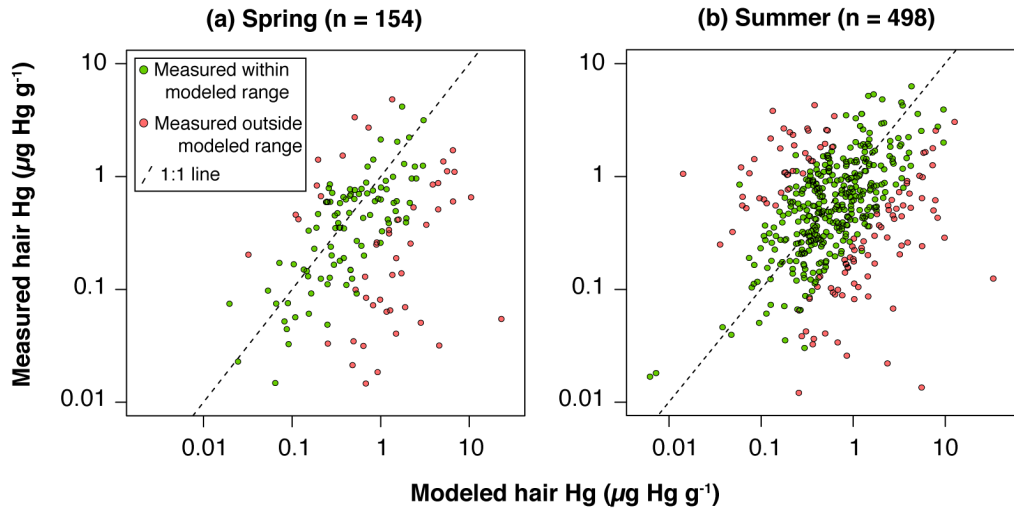


Figure 1. Comparison of measured and probabilistically modeled hair Hg concentrations for Labrador Inuit in the Lower Lake Melville Region during Spring and Summer survey periods. Green circles indicate measured values that fall within the modeled range. Red circles indicate measured values that fall outside the probabilistically modeled range of hair Hg concentrations. Individuals who did not report consuming seafood, birds or marine mammals (8 in Spring, median hair Hg = $0.036 \mu\text{g g}^{-1}$ and 26 in Summer, median hair Hg = $0.049 \mu\text{g g}^{-1}$) are excluded. $R^2 = 0.13$ (Spring), 0.11 (Summer).

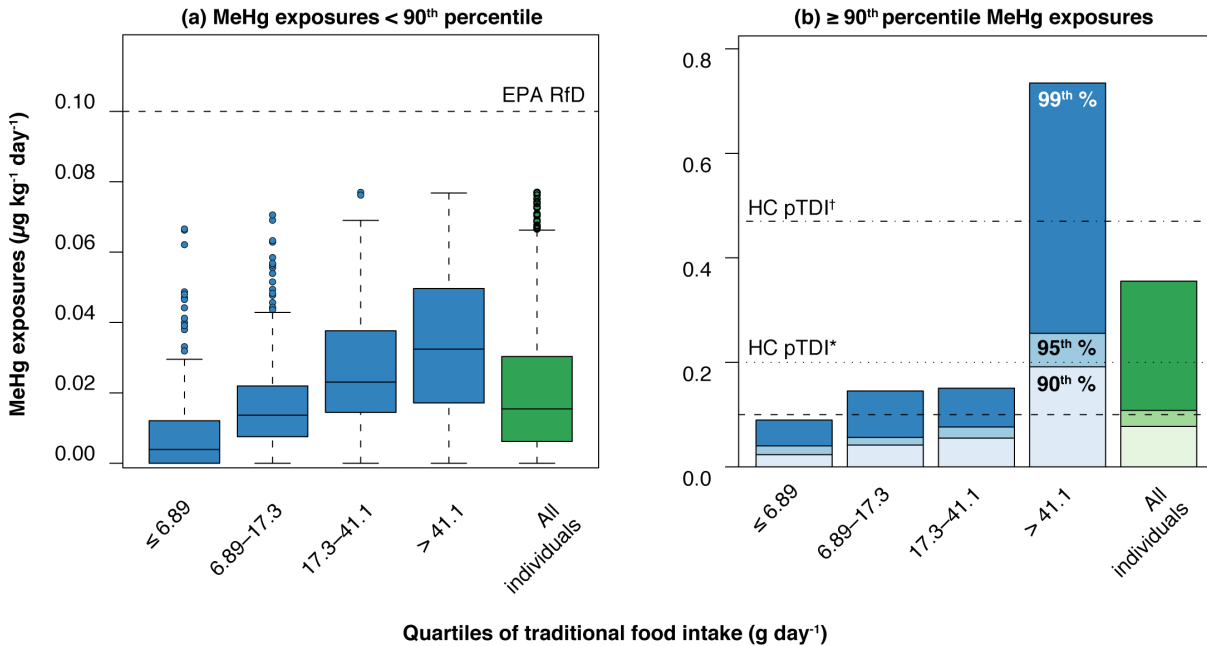


Figure 2. Distributions of MeHg exposures for Inuit in the Lower Lake Melville Region of Labrador by quartile of traditional food consumption (Summer 2014). Panel (a) shows values below the 90th percentile of MeHg exposures and Panel (b) shows the distribution for highly exposed individuals in the population (at or above 90th percentile of MeHg exposures). EPA RfD denotes the U.S. EPA reference dose for methylmercury and HC pTDI indicates the Health Canada provisional tolerable daily intakes for women of childbearing age and children (*) and for everyone else (†).

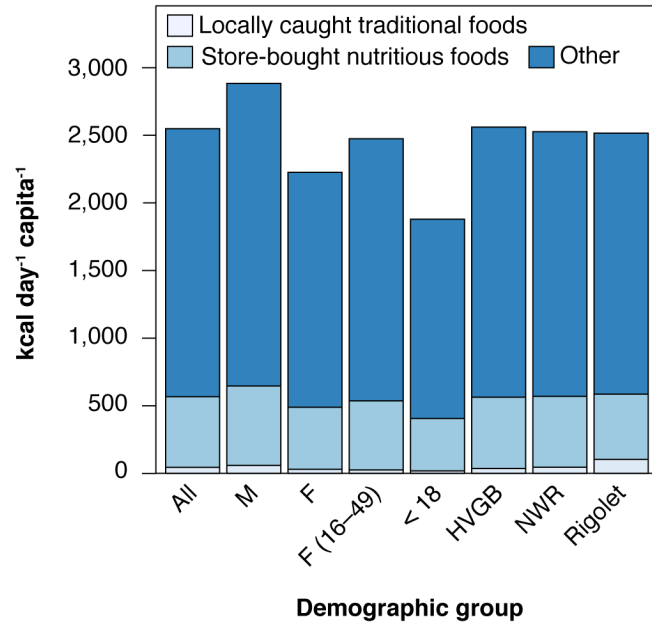


Figure 3. Estimated contributions of different food types to total calories consumed by Inuit in the Lower Lake Melville region of Labrador, Canada. M = male; F = female; HVGB = Happy Valley – Goose Bay; NWR = North West River.

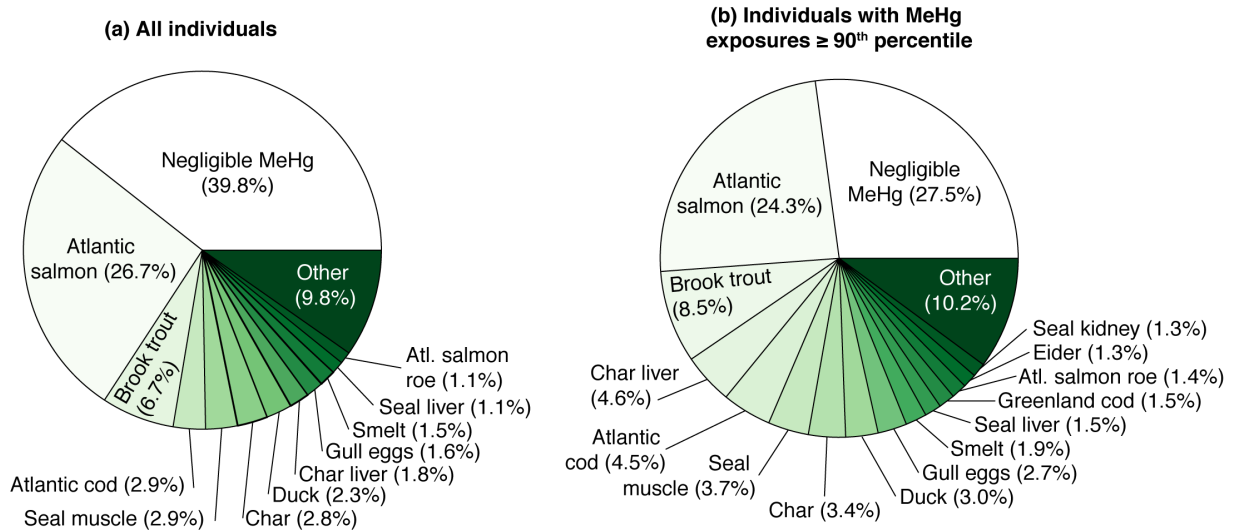


Figure 4. Fraction of caloric intake from traditional foods for top 90% of foods among Lake Melville Inuit for (a) all individuals and (b) individuals with MeHg exposures above the 90th percentile in summer 2014 ($0.08 \mu\text{g kg}^{-1} \text{day}^{-1}$). Dietary data from full-scale summer survey. Widely consumed foods (in the top 90% of contributors to overall calories from traditional foods) include berries, goose, partridge, moose, caribou, seal blubber and rabbit.

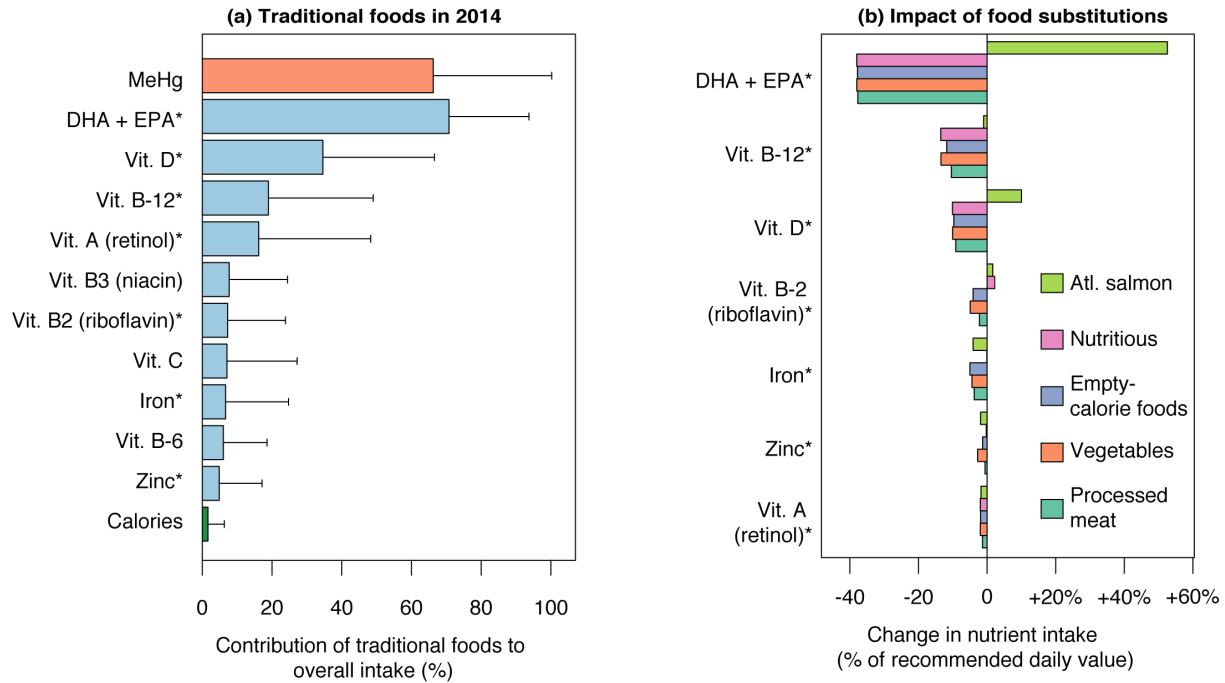


Figure 5. Role of traditional foods for nutrient and MeHg intake and impacts of traditional food substitution. Panel (a) shows the estimated proportion of MeHg and several key nutrients from traditional foods based on survey data for 2014. Panel (b) shows the modeled impact of traditional food substitution on intakes relative to recommended daily values assuming several hypothetical traditional food replacement scenarios. Substitution scenarios for traditional foods include: (1) locally caught Atlantic salmon, (2) nutrient-dense store-bought foods (“Nutritious”), (3) nutrient-sparse junk foods, (4) vegetables and (5) processed meat. Shaded bars represent per-capita averages, and lines represent the 5th–95th percentile individuals. * denotes per-capita intake below recommended daily values based on US FDA recommendations⁹⁴ and 500 mg day⁻¹ for DHA + EPA⁹³.

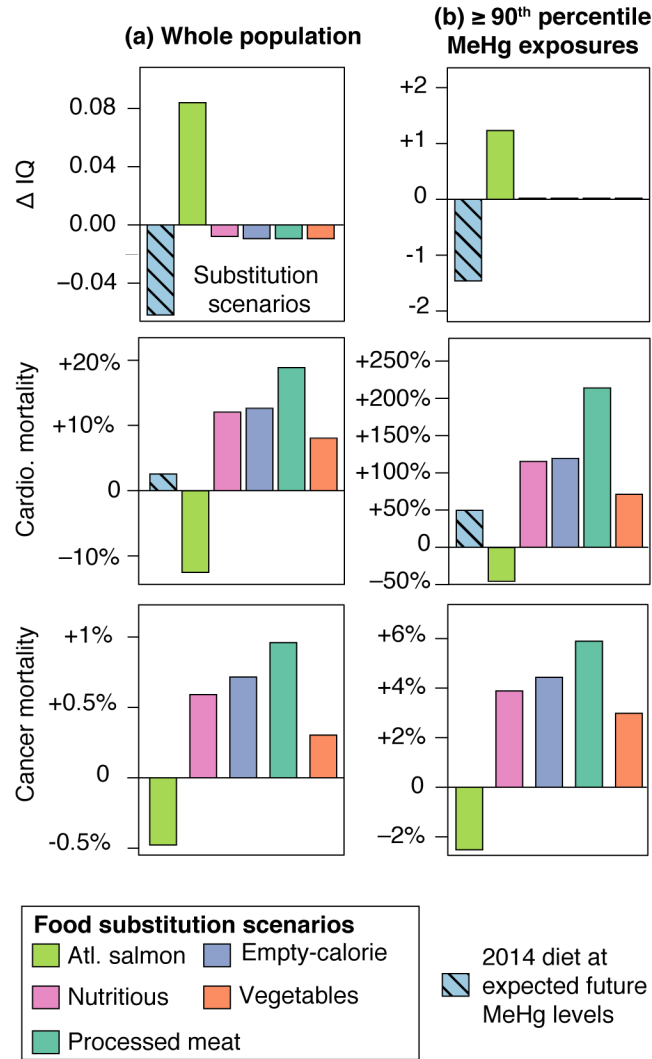


Figure 6. Comparison of neurodevelopmental, cardiovascular and cancer risks for traditional food substitution scenarios compared to risks associated with projected future MeHg levels in traditional foods from Calder et al.⁴² Panel (a) shows median risks for the whole population. Panel (b) shows individuals at or above the 90th percentile of MeHg exposures in summer 2014 ($0.08 \mu\text{g kg}^{-1} \text{day}^{-1}$). Traditional food substitution scenarios include: (1) locally caught Atlantic salmon, (2) the representative basket of subsidized nutrient-dense store-bought foods (“Nutritious”), (3) nutrient-sparse junk foods, (4) vegetables and (5) processed meat. Excess cardiovascular and cancer mortality risks are presented as fractions of present-day risks (+0% corresponds to a relative risk of 1, meaning no change relative to 2014).

Supporting Information

Risk Tradeoffs Associated with Traditional Food Advisories for Labrador Inuit

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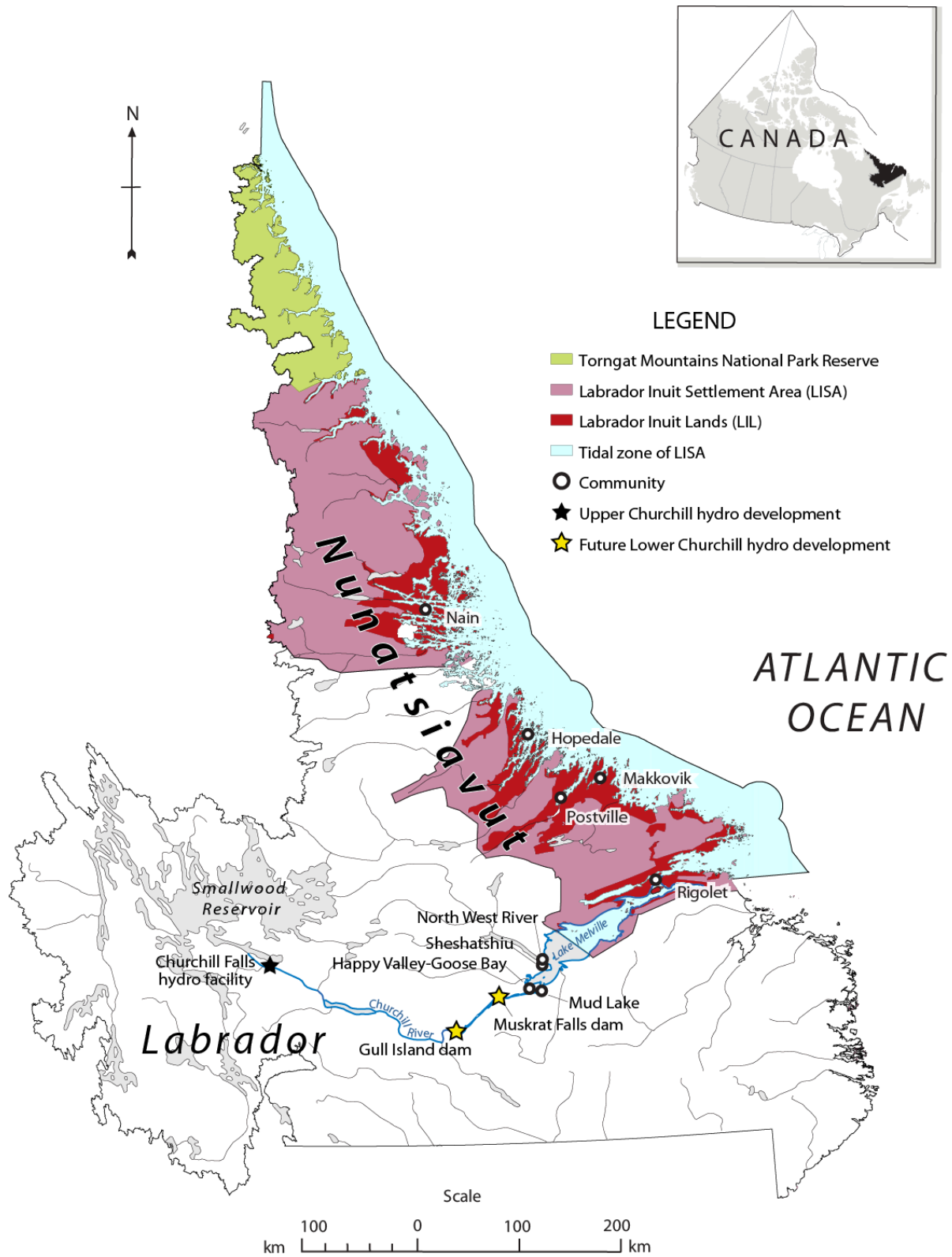


Figure S1: Map of the Labrador Inuit Settlement Area, existing and future hydroelectric developments on the Churchill River, and locations of indigenous communities. Source: Durkalec et al. (2016). Reprinted with permission from Nunatsiavut Government.

Table S1: Food frequency questionnaire data collected in March/April (winter), June/July (spring) and August/September (summer) 2014, adapted from Calder et al. (2016).

Demographic Group	Winter (<i>n</i>)	Spring (<i>n</i>)	Summer (<i>n</i>)	Total (<i>n</i>)	Unique individuals (fraction of Inuit population) ^a
All individuals	231	294	1054 ^b	1579 ^b	1,145 ^b
Non-Inuit household ^c members	34	49	167	250	188
Inuit individuals	197	245	882	1324	952 (38%)
<u>Communities</u>					
Happy Valley-Goose Bay ^d	170	217	667	1054	745 (31%)
North West River	30	34	158	222	167 (43%)
Rigolet	31	43	229	303	233 (87%)
<u>Demographic Group^e</u>					
Women of childbearing age (16-49) ^f	59	77	278	414	306
Children ≤12 years	55	59	166	280	179
Women of childbearing age (16-49) & children ≤ 12 in Rigolet	15	19	100	134	101
All male >12 years	74	108	387	569	406
All female > 49 years	28	37	191	256	200

^a Data from some individuals are for multiple survey periods. Total Inuit population is based on the 2011 Census and National Household Survey (Statistics Canada 2012; Statistics Canada 2013).

^b Total includes three individuals who did not report Inuit status.

^c Non-Inuit individuals who share a household with a registered Inuit beneficiary identified by the Nunatsiavut Government were included in the survey.

^d Includes the nearby community of Mud Lake (*n* = 22).

^e Combined data for all three communities.

^f As defined by the U.S. National Health and Nutrition Examination Survey (McDowell et al. 2004).

Table S2: Demographic breakdown of study participants in summer survey round vs. general population.

	Happy Valley-Goose Bay		North West River		Rigolet	
<i>Male</i>	Sample ^a	Population ^b	Sample ^a	Population ^b	Sample ^a	Population ^b
0–17	64	830	5	60	29	35
18–44	105	1,405	13	85	41	65
45–64	75	1,100	28	95	31	40
≥65	31	320	29	35	10	20
<i>Female</i>						
0–17	80	218	7	50	34	35
18–44	171	407	16	85	45	55
45–64	94	278	31	95	30	40
≥65	41	87	27	35	8	5

^a Summer (August-September) survey round only. Excludes nine participants whose age was not recorded.

^b Age bracket counts available for total (Inuit and non-Inuit) population only rounded to multiples of 5 (Statistics Canada 2013). Sample as a fraction of total Inuit population reported in Table S1.

Table S3: MeHg concentrations, dietary database information and per-capita caloric significance for locally caught traditional seafood

Food name (nutritional category) ^a	Name in nutritional database	Database code ^b	Baseline calories (%) ^c	MeHg Mean ± SD (ng g ⁻¹ ww) ^d
Arctic char liver (FL)	n/a		4.7	75.5
Arctic char muscle (FM)	Fish, arctic char, native, meat, raw	3230	4.9	62.4 ± 41.8
Arctic char roe (FR)	Fish, roe, mixed species, raw	3045	0.1	7.49
Atlantic cod muscle (FM)	Fish, cod (scrod), Atlantic, baked or broiled	3195	5.1	186 ± 57.4
Atlantic salmon liver (FL)	Fish, salmon, native, king or chinook, liver	5899	0.7	88.6
Atlantic salmon muscle (FM)	Fish, salmon, Atlantic, wild, baked or broiled	3156	46.9	73.2 ± 20.1
Atlantic salmon roe (FR)	Fish, salmon, native, eggs, raw	5928	1.9	8.78
Brook trout liver (FL)	n/a		0.8	95.2 ± 34.4
Brook trout muscle (FM)	Fish, trout, brook, raw	7234	11.7	105 ± 34.4
Brook trout roe (FR)	Fish, roe, mixed species, raw	3045	0.6	52.7 ± 22.4
Capelin muscle (FM)	Fish, smelt, rainbow (American, capelin), baked or broiled	3065	0.9	17 ± 2.24
Capelin roe (FR)	Fish, roe, mixed species, raw	3045	0.04	2.04
Clams (SF)	Mollusks, clam, mixed species, boiled or steamed	3111	0.2	10 ± 2
Flatfish (FM)	Fish, flatfish (flounder or sole or plaice), baked or broiled	3007	0.03	68 ± 41
Itiks (SF)	n/a		0.03	40
Lake trout (FM)	Fish, trout, brook, raw	7234	0.9	99 ± 46
Mussels (SF)	Mollusks, mussel, blue, boiled or steamed	3116	1.5	3.46 ± 0.5
Ouananiche (FM)	Fish, salmon, Atlantic, wild, baked or broiled	3156	0.02	150 ± 110
Porpoise blubber (MM)	n/a		0.02	~0
Porpoise liver (MM)	n/a		0	1,220 ± 870
Porpoise meat (MM)	n/a		0.02	600 ± 450
Rock cod liver (FL)	Fish, lingcod, native, liver	5888	1.0	225
Rock cod muscle (FM)	Fish, cod (scrod), Atlantic, baked or broiled	3195	1.5	186 ± 57.4
Scallops (SF)	Mollusks, scallop (bay and sea), cooked, steamed	5634	0.9	22
Sculpin liver (FL)	n/a		0	87.5 ± 86.4
Sculpin muscle (FM)	Fish, sculpin, native, raw	5919	0.1	231 ± 91.8
Seal blubber (MM)	Game meat, native, ringed seal, blubber, boiled	5781	4.1	~0
Seal liver (MM)	Game meat, native, ringed seal, liver, raw	5788	2.0	175 ± 119
Seal kidney (MM)	n/a		1.6	261 ± 74.5

^a FL = fish liver; FM = fish muscle; FR = fish roe; MM = marine mammal; SF = shellfish.

^b From Canada Nutrient File.

^c Fraction of total per-capita calories from all locally caught traditional seafood (summer 2014).

^d From Calder et al. (2016).

n/a: Data not available and values are calculated as average of other foods in same nutritional category.

Table S3 (cont'd): MeHg concentrations, dietary database information and per-capita caloric significance for locally caught traditional seafood

Food name (nutritional category) ^a	Name in nutritional database	Database code ^b	Baseline calories (%) ^c	MeHg Mean \pm SD (ng g ⁻¹ ww) ^d
Seal muscle (MM)	Game meat, native, ringed seal, meat, boiled	5783	5.1	172 \pm 110
Smelt (FM)	Fish, smelt, rainbow (American, capelin), baked or broiled	3065	2.7	114 \pm 49.2
Whale blubber (MM)	Whale, bowhead, skin and subcutaneous fat (muktuk)	35086*	0.06	~0
Whale muscle (MM)	Game meat, whale, raw	3648	0	75 \pm 21
Wrinkles (SF)	Pacific surf, cooked periwinkle meat	45002474*	0.1	40

^a FM = fish muscle; MM = marine mammal; SF = shellfish.

^b Codes with * are from USDA Nutrient Database; other foods are from the Canada Nutrient File.

^c Fraction of total per-capita calories from all locally caught traditional seafood (summer 2014).

^d From Calder et al. (2016).

Table S4: MeHg concentrations, dietary database information and per-capita caloric significance for other locally caught traditional foods

Food name (nutritional category) ^a	Name in nutritional database	Database code ^b	Baseline calories (%) ^c	MeHg Mean ± SD (ng g ⁻¹ ww) ^d
Bakeapples (FR)	Cloudberry (bakeapple), native	5939	9.42	~0
Black bear (TM)	Game meat, native, bear, simmered	3566	0.9	~0
Caribou (TM)	Game meat, native, caribou (reindeer), meat, cooked	3578	7.7	~0
Duck eggs (EG)	Egg, duck, whole, fresh, raw	1138*	1.7	30 ± 3
Duck muscle (FP)	Duck, wild, native, cooked	5931	5.4	117 ± 75.7
Eider muscle (FP)	n/a		2.3	113 ± 33
Goose muscle (FP)	Goose, domesticated, meat only, roasted	672	12.5	~0
Guillemot eggs (EG)	n/a		0	210 ± 9.59
Guillemot muscle (FP)	n/a		0.17	270 ± 70
Gull eggs (EG)	n/a		3.6	59.5 ± 7.84
Gull muscle (FP)	n/a		0.01	230 ± 27.1
Labrador tea (FR)	n/a		0.06	~0
Loon eggs (EG)	n/a		0.03	900 ± 1,880
Loon muscle (FP)	n/a		0	846 ± 237
Moose (TM)	Game meat, native, moose, roasted	3588	9.2	~0
Okalik/Hare (TM)	Game meat, native, rabbit, wild, cooked	3596	0.4	~0
Owl muscle (FP)	n/a		0.01	~0
Partridge muscle (FP)	Spruce grouse, native, meat, cooked		10.5	~0
Polar bear meat (TM)	Game meat, native, polar bear, meat, boiled	5834	0.03	70 ± 50
Rabbit (TM)	Game meat, native, rabbit, wild, cooked	3596	2.3	~0
Red berries (FR)	n/a		12	~0
Sandpiper (FP)	n/a		0.4	70 ± 7
Snowbird (FP)	n/a		0	~0
Tern eggs (EG)	n/a		1.2	424 ± 107
Tern muscle (FP)	n/a		0.23	233 ± 246
Wild blackberries (FR)	Blackberry, raw		1.1	~0
Wild blueberries (FR)	Blueberry, raw		10.2	~0
Wild raspberries (FR)	Raspberry, wild, raw		5.6	~0
Other picked berries (FR)	n/a		2.2	~0
Other wild plants (FR)	n/a		0.6	~0

^a EG = egg; FR = fruit; FP = fowl/poultry; TM = terrestrial mammal; FR = fruit.

^b Codes with * are from USDA Nutrient Database; other foods are from the Canada Nutrient File.

^c Fraction of total per-capita calories from all other locally caught food (summer 2014).

^d From Calder et al. (2016).

n/a: Data not available and values are calculated as average of other foods in same nutritional category.

Table S5: MeHg concentrations, dietary database information and per-capita caloric significance for store-bought seafood

Food name (nutritional category) ^a	Name in nutritional database	Database code ^b	Baseline calories (%) ^c	MeHg Mean ± SD (ng g ⁻¹ ww) ^d
Battered cod (FM)	Sea Cuisine, breaded cod tender flaky fillets	4511– 9629*	11.5	110.0 ± 64.7
Battered haddock (FM)	Gorton's, haddock breaded fish sticks	4504– 5406*	0.9	59.1 ± 26.9
Brook trout (FM)	Fish, trout, brook, raw	7234	0.6	87.9 ± 41.0
Canned oysters (SF)	Mollusks, oyster, eastern (blue point), wild, canned, solids and liquid	3121	0.7	2.6 ± 2.9
Canned salmon (FM)	Fish, salmon, chum (keta), canned, drained, solids with bone, salted	3218	3.7	40.0 ± 19.5
Canned tuna (FM)	Fish, tuna, light, canned in water, drained, unsalted	3131	13.1	162.4 ± 136.8
Catfish (FM)	Fish, catfish (wolffish), Atlantic, baked or broiled	3170	0	40.1 ± 19.0
Clams (SF)	Mollusks, clam, mixed species, boiled or steamed	3111	1.5	10.0 ± 2.0
Crab (SF)	Crustaceans, crab, red, steamed	3238	3.1	60.1 ± 27.0
Fish sticks (pollock) (FM)	Fish, fish sticks, frozen, prepared	3006	3.7	18.7 ± 10.9
Fresh cod (FM)	Fish, cod (scrod), Atlantic, baked or broiled	3195	18.3	110.9 ± 66.0
Fresh pollock (FM)	Fish, pollock, Atlantic (Boston blue), baked or broiled	3152	1.5	19.0 ± 10.6
Fresh tuna (FM)	Fish, tuna, skipjack (aku), baked or broiled	3166	1.2	440.1 ± 246.1
Herring (FM)	Fish, herring, Atlantic, baked or broiled	3015	3.0	18.0 ± 10.2
Lobster (SF)	Crustaceans, lobster, American (northern), boiled or steamed	3210	3.4	36.0 ± 16.9
Mussels (SF)	Mollusks, mussel, blue, boiled or steamed	3116	8.7	24.0 ± 12.5
Rainbow trout (FM)	Fish, trout, rainbow, farmed, baked or broiled	3187	0.9	34.0 ± 17.5
Salmon (FM)	Fish, salmon, Atlantic, wild, baked or broiled	3156	8.7	42.0 ± 19.8
Sardines (FM)	Fish, sardine, Atlantic, canned in oil, drained solids with bone	3203	5.0	34.9 ± 17.5
Scallops (SF)	Mollusks, scallop (bay and sea), cooked, steamed	5634	4.7	22.1 ± 12.1
Shrimp (SF)	Crustaceans, shrimp, mixed species, boiled or steamed	3212	5.9	30.9 ± 15.7
Skate (FM)	n/a		0	118 ± 50.7
Sole (FM)	Fish, flatfish (flounder or sole or plaice), baked or broiled	3007	0.05	101.3 ± 39.8
Tilapia (FM)	Fish, tilapia, baked or broiled	5697	0.1	20.1 ± 10.8

^a FM = fish muscle; SF = shellfish.

^b Codes with * are from USDA Nutrient Database. Other foods are from the Canada Nutrient File.

^c Fraction of total per-capita calories from all store-bought seafood (summer 2014).

^d From Calder et al. (2016).

n/a: Data not available and values are calculated as average of other foods in same nutritional category.

Supplemental methods for dietary survey

Food frequency questionnaire design

We designed the food frequency questionnaire (FFQ) in collaboration with a committee of Inuit elders convened by the Nunatsiavut Government (NG). Through this collaborative process, we developed an exhaustive list of traditional foods eaten by Labrador Inuit. We also explored the feasibility of different ways of measuring food intake (recall, journals, etc.). It was decided that a self-reported recall instrument administered by trained interviewers was likely to maximize enrollment and consistency of results.

We designed standardized clay models for participants to use to describe their average serving size of each food they reported eating. These models corresponded to serving sizes ranging between approximately 115 and 290 g.

Participant recruitment

The NG enrolled participants in the study through bilingual (English/Inuktitut) informational posters dispersed through the communities, informational sessions organized by staff from the NG Environmental Division and informational phone calls to community members registered as NG ‘beneficiaries’ (Inuit with a demonstrated connection to the Labrador Inuit Land Claims Area). Participants were entered into two raffles for three prizes each ranging from 250 to 1,000 CAD to incentivize 1) participation in all three survey periods ($n = 147$) and 2) participating in the larger-scale summer period ($n = 1,054$).

Hair Hg analysis

Trained research assistants collected hair samples from all willing participants in the Spring and Summer survey periods. Hair was collected from the occipital region of the scalp and we analyzed the 2 cm proximal ends of each sample for total Hg as an indicator of MeHg exposure over the past two to three months. The analytical procedure for total mercury used thermal decomposition, amalgamation, and atomic absorption spectrophotometry following US EPA method 7473 on a Nippon MA-3000 or Milestone DMA-80 at Harvard University. One method blank and one sample containing certified hair reference material were tested every 10 samples and all recoveries were within certified ranges. The mean relative standard deviation of replicate samples was 8.6%.

Table S6: Dietary database information and per-capita caloric significance for store-bought nutritious foods

Food name (nutritional category) ^a	Name in nutritional database	Database code ^b	Baseline calories (%) ^c	Retail and consumer waste (%) ^d
All fresh fruits (FR)	Apple, Gala, raw, with skin	7216	6.1	40
All frozen fruits (FR)	Blueberry, frozen, unsweetened	1706	0.04	0
All fresh vegetables, except whole pumpkins (VE)	Brussels sprouts, boiled, drained	2379	4.8	40
All frozen vegetables (excluding French fries, etc.) (VE)	Broccoli, frozen, chopped, boiled, drained	2377	0.08	0
All-purpose flour, whole-wheat flour, rye flour and other semi-perishable flours (except cake flour and pastry flour) (GR)	Grains, wheat flour, white, all purpose, bleached	4501	0	29
Bacon (PM)	Pork, cured, back bacon, pan-fried	7219	4.8	15
Bread (except garlic bread) (GR)	Bread, white, commercial	4066	8.3	29
Bread products without filling or coating (GR)	English muffin, wheat	3906	2.9	29
Butter (DA)	Butter, regular	118	1.4	15.5
Cheese (including block cheese, shredded cheese and cottage cheese) (DA)	Cheese, cheddar	119	1.2	15.5
Chocolate milk (DA)	Milk, fluid, chocolate, whole	69	1.7	15.5
Combination foods (e.g., lasagna)	Lasagna with meat and sauce, frozen	5870	0.5	0
Cook-type cereals (e.g., oatmeal and porridge) (GR)	Cereal, hot, oats, porridge	1432	0.03	29
Cooking oils (e.g., canola, peanut, olive and linseed)	Vegetable oil, canola	451	1.0	0
Crackers, crisp bread, hard bread, Pilot biscuits, Melba toast, arrow-root biscuits and Social Tea biscuits (GR)	Snacks, rice cakes, crackers (include mini rice cakes)	5493	1.1	29

^a Food name/category from the Nutrition North program; FM = FR = fruit; VE = vegetables; GR = grains; PM = processed meat; DA = dairy.

^b From the Canada Nutrient File.

^c Fraction of total per-capita calories from all store-bought nutritious foods (summer 2014).

^d From Gustavsson et al. (2011).

Table S6 (cont'd): MeHg concentrations, dietary database information and per-capita caloric significance for store-bought nutritious foods

Food name (nutritional category) ^a	Name in nutritional database	Database code ^b	Baseline calories (%) ^c	Retail and consumer waste (%) ^d
Cream, sour cream, cream cheese (DA)	Cream, sour, cultured, 18% M.F.	152	1.2	15.5
Dried fruits (e.g., grapes, dates, cranberries and apricots) (FR)	Apple, dried, sulphured, uncooked	1490	0.2	67.2
Eggs and egg substitutes (EG)	Egg, chicken, whole, fresh or frozen, raw	125	4.4	15
Enriched soy milk	Plant-based beverage, soy, enriched, all flavours	6720	0.2	15.5
Fresh and frozen meat other than side bacon and products that are breaded, battered or in pastry (RM)	Beef, ground, lean, raw	2683	21.7	15
Fresh and frozen pasta (except combined foods containing pasta) (GR)	Pasta, fresh-refrigerated, plain, as purchased	4502	0	29
Fresh and frozen pizzas	Fast foods, pizza, cheese, meat and vegetable, regular crust, frozen, cooked	5862	4.9	0
Fresh and frozen poultry e.g., chicken, turkey, goose) other than products that are breaded, battered or in pastry (FP)	Chicken, broiler, breast, skinless, boneless, meat, grilled	7322	5.8	15
Fresh and frozen seafood, other than products that are breaded, battered or in a pie crust (FM)	Fish, tuna, white, canned with water, drained, salted	3084	Counted in Table S4	n/a
Fresh milk (DA)	Milk, fluid, partly skimmed, 2% M.F.	61	2.6 ^e	15.5
Frozen French fries, home fries and similar potato-based products	Potato, french-fried, frozen, shoestring, heated in oven	6517	5.3	0
Ice cream, iced milk, iced yogourt and sorbet (DA)	Dessert, frozen, ice cream, chocolate	4288	2.6 ^g	15.5
Individually wrapped unsweetened juice (all sizes)	Apple juice, canned or bottled, unsweetened, calcium and Vitamin C and D added	7419	0.8	0
Lard and shortening	Shortening, household, unspecified vegetable and animal oils	539	1.5	0

^a Food name/category from the Nutrition North program; FM = FR = fruit; VE = vegetables; GR = grains; PM = processed meat; DA = dairy.

^b From the Canada Nutrient File.

^c Fraction of total per-capita calories from all store-bought nutritious foods (summer 2014).

^d From Gustavsson et al. (2011).

Table S6 (cont'd): MeHg concentrations, dietary database information and per-capita caloric significance for store-bought nutritious foods

Food name (nutritional category) ^a	Name in nutritional database	Database code ^b	Baseline calories (%) ^c	Retail and consumer waste (%) ^d
Margarine	Margarine, tub, hydrogenated, canola oil	7575	1.5	0
Melted cheese spreads (e.g., Cheez Whiz)	Kraft Cheez Whiz Pasteurized Process Cheese Sauce	1188*	0.04	0
Peanut butter and other nut butters	Peanut butter, natural	6289	0.7	81.4
Perishable dips	Dip, cream cheese base	6786	0.5 ^f	0
Powdered and evaporated milk	Milk, evaporated, whole, canned, undiluted, 7.8% M.F.	140	0.02	0
Processed cheese (e.g., Velveeta)	Kraft Velveeta Pasteurized Process Cheese Spread	1191*	1.8	0
Ready-to-eat breakfast cereals (GR)	Cereal, ready to eat, Life, Quaker	1258	0.8	29
Salad dressing and mayonnaise	Salad dressing, mayonnaise type, commercial, regular	527	1.3	0
Tofu and other vegetable-based meat substitutes (e.g., vegetable patties and nut burgers)	Tofu, silken, firm	4911	0	0
UHT milk (DA)	Milk, fluid, whole, pasteurized, homogenized, 3.25% M.F.	113	1.0 ^h	15.5
Unsweetened nuts and grains (GR)	Nuts, mixed nuts, dry roasted with peanuts	2577	0.6	0
Yogurt and yogurt drinks (DA)	Yogourt, plain (2-3.9% M.F.)	6961	1.3 ^f	15.5

^a Food name/category from the Nutrition North program; GR = grains; DA = dairy.

^b From the Canada Nutrient File.

^c Fraction of total per-capita calories from all store-bought nutritious foods (summer 2014).

^d From Gustavsson et al. (2011).

Table S7: Probability distribution of peak MeHg concentrations in locally caught traditional foods following upstream hydroelectric development, first reported by Calder et al. (2016)

Species	Expected mean	75 th percentile	90 th percentile	95 th percentile
Arctic char				
Muscle	0.41	0.51	0.78	1.0
Liver	0.49	0.58	0.70	0.80
Roe	0.05	0.06	0.07	0.08
Atlantic cod	0.41	0.50	0.65	0.76
Atlantic salmon				
Muscle	0.16	0.20	0.25	0.29
Liver	0.20	0.23	0.28	0.31
Roe	0.020	0.023	0.027	0.031
Black duck				
Muscle	0.44	0.55	0.83	1.1
Eggs	0.11	0.13	0.16	0.18
Brook trout				
Muscle	0.68	0.84	1.1	1.3
Liver	0.62	0.76	1.0	1.2
Roe	0.34	0.42	0.58	0.70
Capelin				
Muscle	0.04	0.05	0.06	0.07
Roe	0.01	0.01	0.01	0.01
Clams	0.03	0.03	0.04	0.04
Eider				
Muscle	0.20	0.24	0.30	0.34
Flatfish	0.17	0.22	0.32	0.40
Green sea urchin	0.10	0.12	0.14	0.16
Guillemot				
Muscle	0.68	0.82	1.0	1.2
Eggs	0.53	0.61	0.74	0.84
Gull				
Muscle	0.41	0.46	0.54	0.59
Eggs	0.15	0.18	0.21	0.24
Lake trout	1.0	1.3	1.8	2.2
Loon				
Eggs	5.6	5.7	13.3	20.9
Minke whale	0.07	0.09	0.10	0.11
Mussels	0.01	0.01	0.01	0.01
Ouananiche	1.5	1.9	3.0	3.9
Periwinkles	0.10	0.12	0.14	0.16
Porpoise				
Muscle	1.4	1.8	2.7	3.5
Liver	2.8	3.6	5.2	6.8
Rock cod				
Muscle	0.42	0.50	0.65	0.77
Liver	0.50	0.58	0.70	0.79

Table S8 (cont'd): Probability distribution of peak MeHg concentrations in locally caught traditional foods following upstream hydroelectric development, first reported by Calder et al. (2016)

Species	Expected mean	75 th percentile	90 th percentile	95 th percentile
Scallops	0.06	0.07	0.08	0.09
Sculpin				
Muscle	0.54	0.66	0.88	1.0
Liver	0.20	0.24	0.42	0.58
Seal				
Muscle	0.66	0.82	1.3	1.6
Liver	0.67	0.84	1.3	1.7
Kidney	1.0	1.2	1.6	1.9
Smelt	0.29	0.36	0.48	0.58
Tern	0.41	0.50	0.86	1.2

Derivation of dose-response functions for cardiovascular and cancer risks

Overall relative risks for cardiovascular and cancer death for each scenario are calculated as the product of individual relative risk equations developed based on dose-response information from the literature. This is presented in Equation S1 where RR is the relative risk for cancer or cardiovascular death (i) for all individual nutrients, foods and contaminants (j) considered and presented in Tables 2 and 3 in the main text.

$$RR_i = \prod_{j=1}^n RR_{ij} \quad [S1]$$

Cause-specific relative risks of cancer or cardiovascular death (i) are calculated as a function of the change in dose (Δ) in nutrient, food or contaminant (j) in model dietary scenarios based on the relative risk calculated using the dose-response information the literature, scaling according to the magnitude of the dose of the substitution in the dietary model vs. the dose considered in the literature (Equation S2).

$$RR_{ij} = \exp\left(\log(RR_{ij}^{lit}) \times \frac{\Delta_j^{model}}{\Delta_j^{lit}}\right) \quad [S2]$$

Dose-response information in the literature is often presented for specific causes of mortality (e.g., relative risk of death from ischemic stroke or from pancreatic cancer) (k). To express this in terms of relative risk of overall mortality (e.g., relative risk of cardiovascular death or cancer death) (i), we consider the fraction at baseline of more general mortality associated with the more specific causes of death (f) for which dose-response information is available, according to Equation S3.

$$RR_{ij}^{lit} = 1 + \left(f_{ik}(RR_{kj}^{lit} - 1)\right) \quad [S3]$$

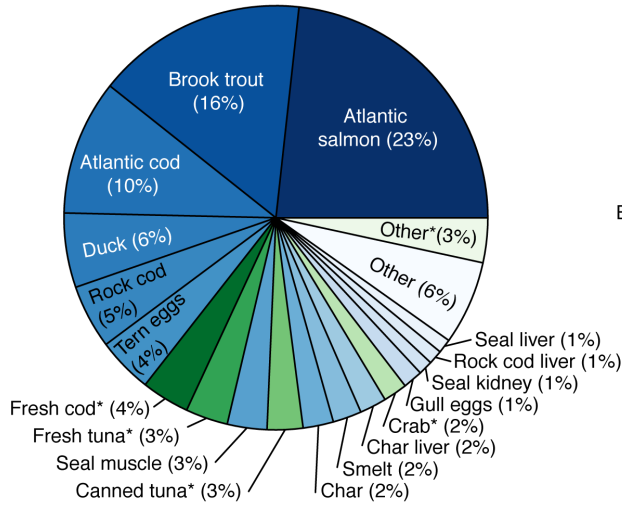
Tables 2 and 3 in the main text present values for RR_{jk}^{lit} , f_{ik} and Δ_j^{lit} for cardiovascular and cancer dose-response functions, respectively.

Table S8: Summary statistics for modeled population-wide MeHg exposures

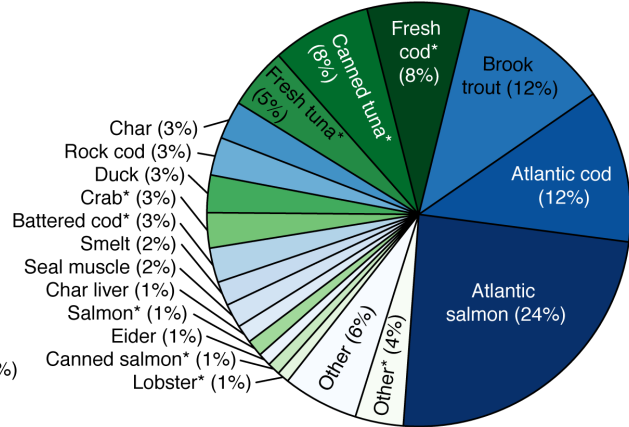
Demographic Group	MeHg exposure percentiles ($\mu\text{g kg}^{-1} \text{ day}^{-1}$)				Fraction of population exceeding threshold	
	50 th (median)	90 th	95 th	99 th	US EPA RfD ($0.1 \mu\text{g kg}^{-1} \text{ day}^{-1}$)	Health Canada pTDI ^a
<i>Winter 2014</i>						
All individuals	0.03	0.11	0.14	0.22	10.45%	1.24%
<i>Spring 2014</i>						
All individuals	0.01	0.05	0.09	0.16	4.29%	0.02%
<i>Summer 2014</i>						
All individuals	0.02	0.08	0.12	0.26	6.85%	0.89%
<u>Communities</u>						
Happy Valley-Goose Bay	0.02	0.07	0.09	0.21	4.88%	0.71%
North West River	0.02	0.07	0.09	0.19	3.90%	0%
Rigolet	0.05	0.2	0.27	0.5	24.38%	3.14%
<u>Males</u>						
All males	0.02	0.1	0.16	0.36	10.35%	1.35%
M 0–17	0.01	0.07	0.21	0.31	7.93%	3.68%
M 18–44	0.02	0.07	0.11	0.21	5.82%	0.15%
M 45–64	0.03	0.12	0.18	0.62	13.40%	1.42%
M ≥ 65	0.04	0.2	0.23	0.37	23.28%	0.36%
<u>Females</u>						
All females	0.02	0.06	0.08	0.14	3.42%	0.44%
F 0–17	0.01	0.06	0.09	0.17	4.70%	0.58%
F 18–44	0.01	0.04	0.06	0.16	2.28%	0.55%
F 45–64	0.02	0.06	0.07	0.13	2.87%	0.31%
F ≥ 65	0.03	0.09	0.1	0.16	6.93%	0%

^a Tolerable daily intake = $0.2 \mu\text{g kg}^{-1} \text{ day}^{-1}$ for women of childbearing age and children 12 and under; $0.47 \mu\text{g kg}^{-1} \text{ day}^{-1}$ for others (Health Canada 2004; Health Canada 2007). Differences in the regulatory thresholds used by these two agencies mainly reflects variability in the uncertainty factor used to account for exposures of sensitive groups.²⁶ During the establishment of the RfD for MeHg by the U.S. EPA in 2000, it was acknowledged that, “It is also important to note that no evidence of a threshold arose for methylmercury-related neurotoxicity within the range of exposures in the Faroe Islands study.”¹⁶ A variety of epidemiological studies have since noted effects associated with exposures below the U.S. EPA’s RfD.^{90,91}

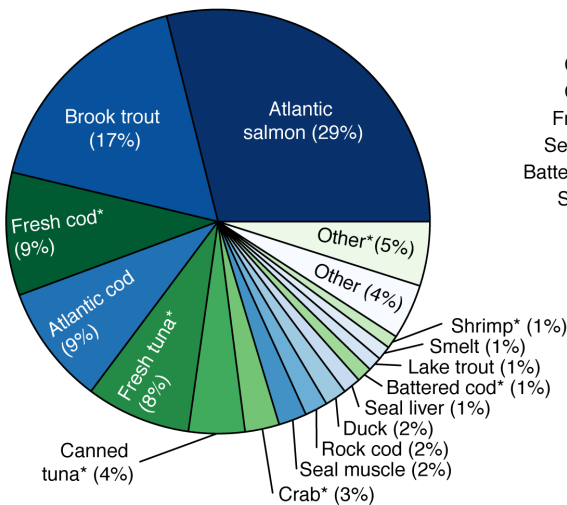
(a) Individuals with MeHg exposures \geq 90th percentile



(b) All individuals – Happy Valley – Goose Bay



(c) All individuals – North West River



(d) All individuals – Rigolet

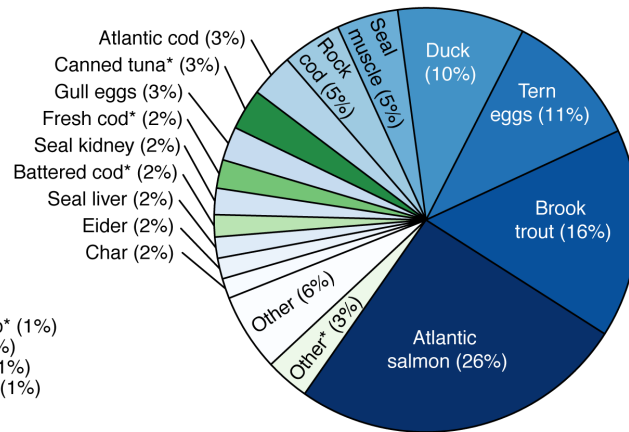


Figure S2: Sources of per-capita MeHg intakes among Lake Melville Inuit for individuals in all three communities with MeHg exposures \geq 90th percentile (a) and for all individuals in each community (b, c, d), Summer 2014. Foods in blue are locally caught. Foods in green with * are store-bought. Rock cod is the local name for *Gadus ogac* (Greenland cod).

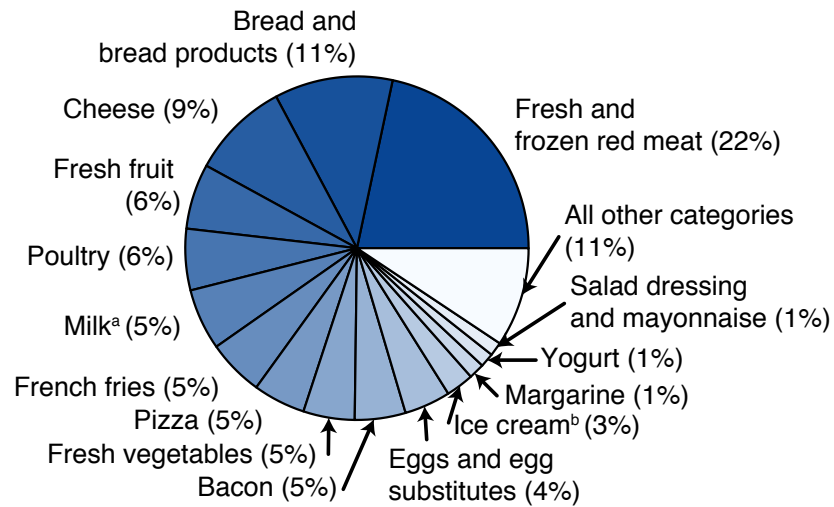


Figure S3: Main categories of store-bought nutritious foods as a fraction of total store-bought nutritious food intake at present day (by calories). ^a Includes soy, evaporated, powdered and chocolate milk. ^b Includes iced yogurt, iced milk and sorbet.

Table S9: Traditional food intake by season and demographic group

Demographic Group	Traditional food intake percentiles (g day ⁻¹)					
	25 th	50 th	75 th	90 th	95 th	99 th
<i>Winter 2014</i>						
All individuals	8.5	21.14	47.68	115.49	160.55	209.96
<i>Spring 2014</i>						
All individuals	5.14	11.62	29.79	77.83	125.95	229.23
<i>Summer 2014</i>						
All individuals	6.86	17.24	41.07	93.22	139.45	276.3
<u>Communities</u>						
Happy Valley-Goose Bay	6.38	14.11	33.5	85.8	121.8	237.25
North West River	10.38	21.96	41.06	87.79	127.48	183.57
Rigolet	16.8	41.32	91.96	180.04	226.56	421.92
<u>Males</u>						
All males	8.63	19.49	50.54	109.5	163.76	410.13
M 0–17	2.24	6.61	19.38	70.39	85.8	309.93
M 18–44	8.63	16.18	36.98	86.53	109.5	194.3
M 45–64	19.1	36.21	80.87	156.55	237.25	434.33
M ≥65	27.48	49.99	107.67	209.35	216.81	306.92
<u>Females</u>						
All females	6.61	14.06	30.28	82.14	110.5	173.09
F 0–17	4.39	6.84	14.19	42.53	82.14	117.07
F 18–44	6.37	11.56	25.42	56.77	99.21	146.18
F 45–64	10.94	20.45	45.02	90.94	130.84	189.08
F ≥65	16.77	32.73	74.48	136.98	196.34	230.9

Table S10: Modeled neurodevelopmental impacts (change in IQ, prenatal exposures, based on Summer 2014)^a

Demographic Group	Baseline diet, increased MeHg	Replacement of traditional foods				
		Nutritious ^b	Empty-calorie	Processed meat	Vegetables	Atl. salmon
All individuals	-0.06 (0.04)	-0.01 (0.02)	-0.01 (0.02)	-0.01 (0.02)	-0.01 (0.02)	0.08 (0.02)
Communities						
Happy Valley-Goose Bay	-0.05 (0.03)	-0.01 (0.01)	-0.01 (0.01)	-0.01 (0.01)	-0.01 (0.01)	0.08 (0.01)
North West River	-0.05 (0.03)	-0.01 (0.01)	-0.01 (0.01)	-0.01 (0.01)	-0.01 (0.01)	0.1 (0.02)
Rigolet	-0.29 (0.18)	-0.04 (0.05)	-0.04 (0.05)	-0.04 (0.05)	-0.04 (0.05)	0.31 (0.06)

^a Values presented are the modeled expected mean (SD) of medians for the whole population and for each community among women of childbearing age (16–49).

^b Representative basket of nutritious store-bought foods (SI Figure S2).

Table S11: Modeled cardiovascular impacts (relative risk of cardiovascular mortality)^a

Demographic Group	Baseline diet, increased MeHg	Replacement of traditional foods with store-bought alternatives				
		Nutritious foods ^b	Empty-calorie	Processed meat	Vegetables	Atl. salmon
All individuals	1.02 (0.01)	1.12 (0.05)	1.12 (0.05)	1.19 (0.06)	1.08 (0.05)	0.88 (0.04)
<u>Communities</u>						
Happy Valley-Goose Bay	1.02 (0.01)	1.1 (0.04)	1.1 (0.04)	1.14 (0.05)	1.07 (0.04)	0.9 (0.04)
North West River	1.03 (0.02)	1.15 (0.07)	1.16 (0.07)	1.25 (0.09)	1.12 (0.07)	0.85 (0.05)
Rigolet	1.12 (0.07)	1.28 (0.14)	1.3 (0.14)	1.54 (0.21)	1.2 (0.14)	0.68 (0.09)
<u>Males</u>						
All males	1.03 (0.02)	1.17 (0.07)	1.17 (0.08)	1.28 (0.1)	1.13 (0.08)	0.82 (0.06)
M 25–44	1.02 (0.01)	1.09 (0.04)	1.1 (0.04)	1.13 (0.04)	1.05 (0.04)	0.88 (0.04)
M 45–64	1.05 (0.02)	1.23 (0.1)	1.23 (0.1)	1.35 (0.12)	1.16 (0.1)	0.81 (0.06)
M ≥65	1.07 (0.05)	1.27 (0.13)	1.29 (0.14)	1.53 (0.21)	1.22 (0.13)	0.74 (0.08)
<u>Females</u>						
All females	1.02 (0.01)	1.08 (0.03)	1.09 (0.04)	1.14 (0.05)	1.06 (0.04)	0.91 (0.03)
F 25–44	1.01 (0)	1.04 (0.01)	1.04 (0.02)	1.06 (0.02)	1.02 (0.01)	0.93 (0.02)
F 45–64	1.02 (0.01)	1.12 (0.05)	1.12 (0.05)	1.18 (0.06)	1.09 (0.05)	0.9 (0.04)
F ≥65	1.05 (0.03)	1.16 (0.07)	1.17 (0.07)	1.29 (0.11)	1.11 (0.07)	0.84 (0.06)

^a Values presented are the modeled expected mean (SD) of medians for the whole population and for each demographic group among individuals at least 25 years of age.

^b Representative basket of nutritious store-bought foods (SI Figure S2).

Table S12: Modeled cancer impacts (relative risk of cancer mortality)^a

Demographic Group	Replacement of traditional foods with store-bought alternatives				
	Nutritious foods ^b	Nutrient-sparse	Processed meat	Vegetables	Atl. salmon
All individuals	1.006 (0.002)	1.007 (0.002)	1.009 (0.003)	1.003 (0.002)	0.995 (0.002)
<u>Communities</u>					
Happy Valley-Goose Bay	1.005 (0.002)	1.006 (0.002)	1.007 (0.002)	1.002 (0.002)	0.996 (0.001)
North West River	1.007 (0.003)	1.009 (0.003)	1.012 (0.003)	1.004 (0.003)	0.995 (0.002)
Rigolet	1.017 (0.007)	1.019 (0.007)	1.024 (0.007)	1.011 (0.007)	0.986 (0.005)
<u>Males</u>					
All males	1.009 (0.003)	1.01 (0.003)	1.013 (0.003)	1.005 (0.004)	0.994 (0.002)
M 25–44	1.004 (0.001)	1.004 (0.002)	1.006 (0.002)	1.001 (0.002)	0.995 (0.002)
M 45–64	1.011 (0.004)	1.013 (0.004)	1.016 (0.005)	1.006 (0.004)	0.992 (0.003)
M ≥65	1.016 (0.006)	1.018 (0.006)	1.023 (0.007)	1.011 (0.006)	0.991 (0.003)
<u>Females</u>					
All females	1.005 (0.002)	1.005 (0.002)	1.007 (0.002)	1.002 (0.002)	0.996 (0.001)
F 25–44	1.002 (0.001)	1.003 (0.001)	1.004 (0.001)	1 (0.001)	0.997 (0.001)
F 45–64	1.006 (0.002)	1.007 (0.002)	1.009 (0.003)	1.003 (0.002)	0.996 (0.002)
F ≥65	1.009 (0.004)	1.011 (0.004)	1.016 (0.004)	1.004 (0.004)	0.995 (0.002)

^a Values presented are the modeled expected mean (SD) of medians for the whole population and for each demographic group among individuals at least 25 years of age. Increased MeHg content in local foods is assumed to have no impact on cancer risks.

^b Representative basket of nutritious store-bought foods (SI Figure S2)

References

- Calder, R. S. D., A. T. Schartup, M. Li, A. P. Valberg, P. H. Balcom and E. M. Sunderland (2016). "Future impacts of hydroelectric power development on methylmercury exposures of Canadian indigenous communities." Environ Sci Technol **50**(23): 13115-22.
- Durkalec, A., T. Sheldon and T. Bell, (Eds.) (2016). Scientific Report. Lake Melville: Avativut, Kanuittailinnivut. Nain, NL: Nunatsiavut Government.
- Gustavsson, J., C. Cederberg, U. Sonersson, R. van Otterdijk and A. Meybeck (2011). "Global Food Losses and Food Waste: Extent, Causes and Prevention." Rome, Italy: United Nations Food and Agriculture Organization (UN FAO).
- Health Canada (2004). "Mercury: Your Health and the Environment: A Resource Tool." Retrieved 2017-11-21 from <https://www.canada.ca/en/health-canada/services/environmental-workplace-health/reports-publications/environmental-contaminants/mercury-your-health-environment-resource-tool.html>.
- Health Canada (2007). "Updating the Existing Risk Management Strategy for Mercury in Retail Fish." Ottawa, ON.
- McDowell, M. A., C. F. Dillon, J. Osterloh, P. M. Bolger, E. Pellizzari, R. Fernando, . . . K. R. Mahaffey (2004). "Hair Mercury Levels in U.S. Children and Women of Childbearing Age: Reference Range Data from NHANES 1999-2000." Environ Health Perspect **112**(11): 1165-71.
- Statistics Canada (2012). "Census Profile." 2011 Census. Retrieved 2018-03-18 from <http://www12.statcan.gc.ca/census-recensement/2011/dp-pd/prof/index.cfm?Lang=E>.
- Statistics Canada (2013). "2011 National Household Survey." Retrieved 2018-03-18 from <http://www12.statcan.gc.ca/nhs-enm/2011/dp-pd/prof/index.cfm?Lang=E>.